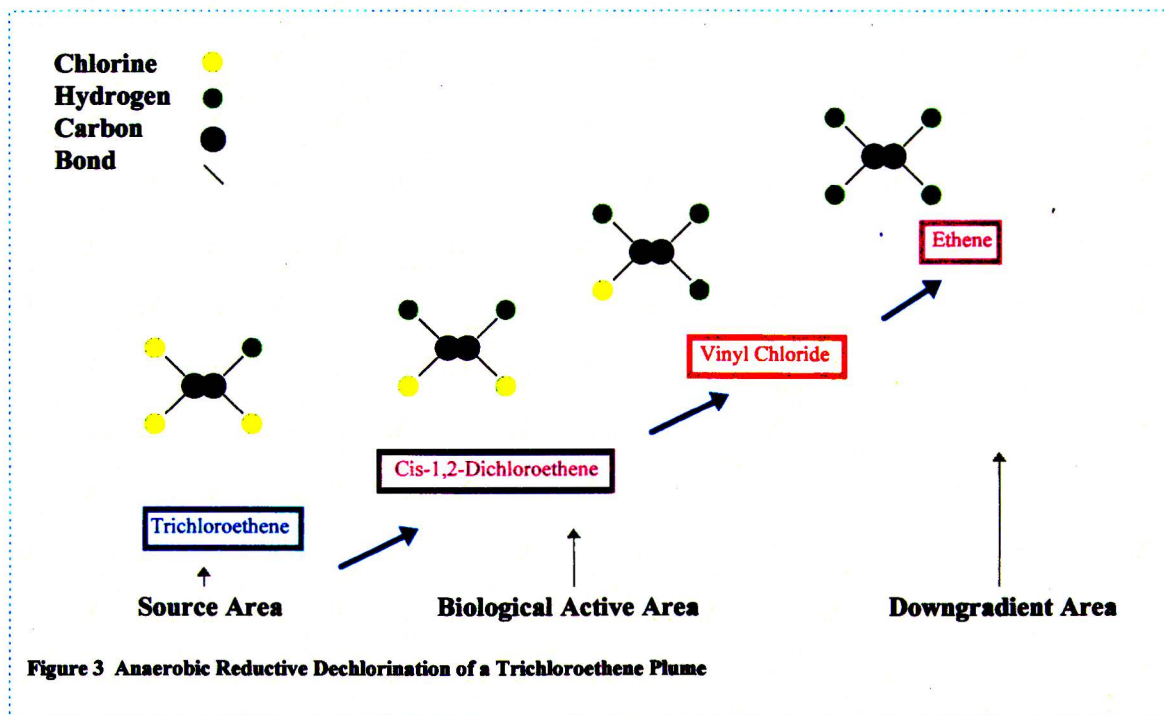


# CONSIDERATIONS FOR THE COLLECTION AND EVALUATION OF NATURAL ATTENUATION PARAMETERS AT SITES CONTAMINATED WITH CHLORINATED SOLVENTS



Prepared by Florida Department of  
Environmental Protection (FDEP)  
Division of Waste Management  
Bureau of Waste Cleanup  
Hazardous Waste Cleanup Section  
April 1999

## TABLE OF CONTENTS

SECTION	TITLE	PAGE
1.0	INTRODUCTION.....	1
2.0	GENERAL OVERVIEW OF CHLORINATED ALIPHATIC HYDROCARBON BIODEGRADATION.....	2
	2.1 Bacteria	
	2.2 Electron Acceptor Reactions	
	2.3 Electron Donor Reactions	
	2.4 Co-Metabolism	
3.0	NATURAL ATTENUATION BY PHYSICAL PROCESSES.....	4
4.0	METABOLIC BY-PRODUCTS AND REDOX PROCESSES IN GROUNDWATER SYSTEMS.....	4
5.0	EVIDENCE USED TO SUPPORT NATURAL ATTENUATION.....	5
	5.1 Direct Evidence	
	5.2 Examine Changes	
	5.3 Laboratory microcosm studies	
6.0	METHODOLOGY USED FOR DATA REVIEW AND SITE EVALUATION.....	6
7.0	NATURAL ATTENUATION SAMPLING AND DATA EVALUATION METHODOLOGY.....	9
	7.1 Monitor Well Placement and Sampling	
	7.2 Evaluation of Biological Mediated Natural Attenuation Data	
8.0	MONITORING THE EFFECTIVENESS OF NATURAL ATTENUATION.....	11
	9.1 Estimation of Biodegradation Rate	
	9.2 Estimation of Overall Natural Attenuation Rate	
9.0	MONITORED NATURAL ATTENUATION REPORTS.....	12
	REFERENCES.....	14

## **Figures**

- Figure 1:** Natural Attenuation of Chlorinated Solvents in Groundwater  
**Figure 2:** Common Degradation Pathways  
**Figure 3:** Anaerobic Reductive Dechlorination of a Trichloroethene Plume  
**Figure 4:** Reductive Dechlorination (Chemical Sequence Example)  
**Figure 5:** Common Patterns of Chlorinated Solvent Biodegradation in a Sequential Aerobic/Anaerobic System  
**Figure 6:** Redox Potential

## **Tables**

- Table 1:** Geochemical Parameters  
**Table 2:** Contaminants and Daughter Products  
**Table 3:** Data Elements  
**Table 4:** Natural Attenuation Parameters and Test Methods  
**Table 5:** Analytical Parameters and Data Evaluation

## **LIST OF APPENDICES**

### **Appendix A**

### **GLOSSARY OF TERMS**

**Please direct all comments and questions concerning this document:**

FDEP

Judie A. Kean, Project Manager, HWC

MS 4520

2600 Blairstone Road

Tallahassee, Florida 32399

850-488-0190.

# **CONSIDERATIONS FOR THE COLLECTION AND EVALUATION OF NATURAL ATTENUATION PARAMETERS AT SITES CONTAMINATED WITH CHLORINATED SOLVENTS**

**(This document is an aggregation of technical information, compiled from professional journals and presentations on natural attenuation. It has been prepared by the Hazardous Waste Cleanup Section of the Bureau of Waste Cleanup as a resource for the Department staff and Department contractors for evaluating and assessing drycleaning solvent contamination sites. This document is not intended to serve as a regulatory standard or to supercede any specific regulatory requirements.)**

## **1.0 INTRODUCTION**

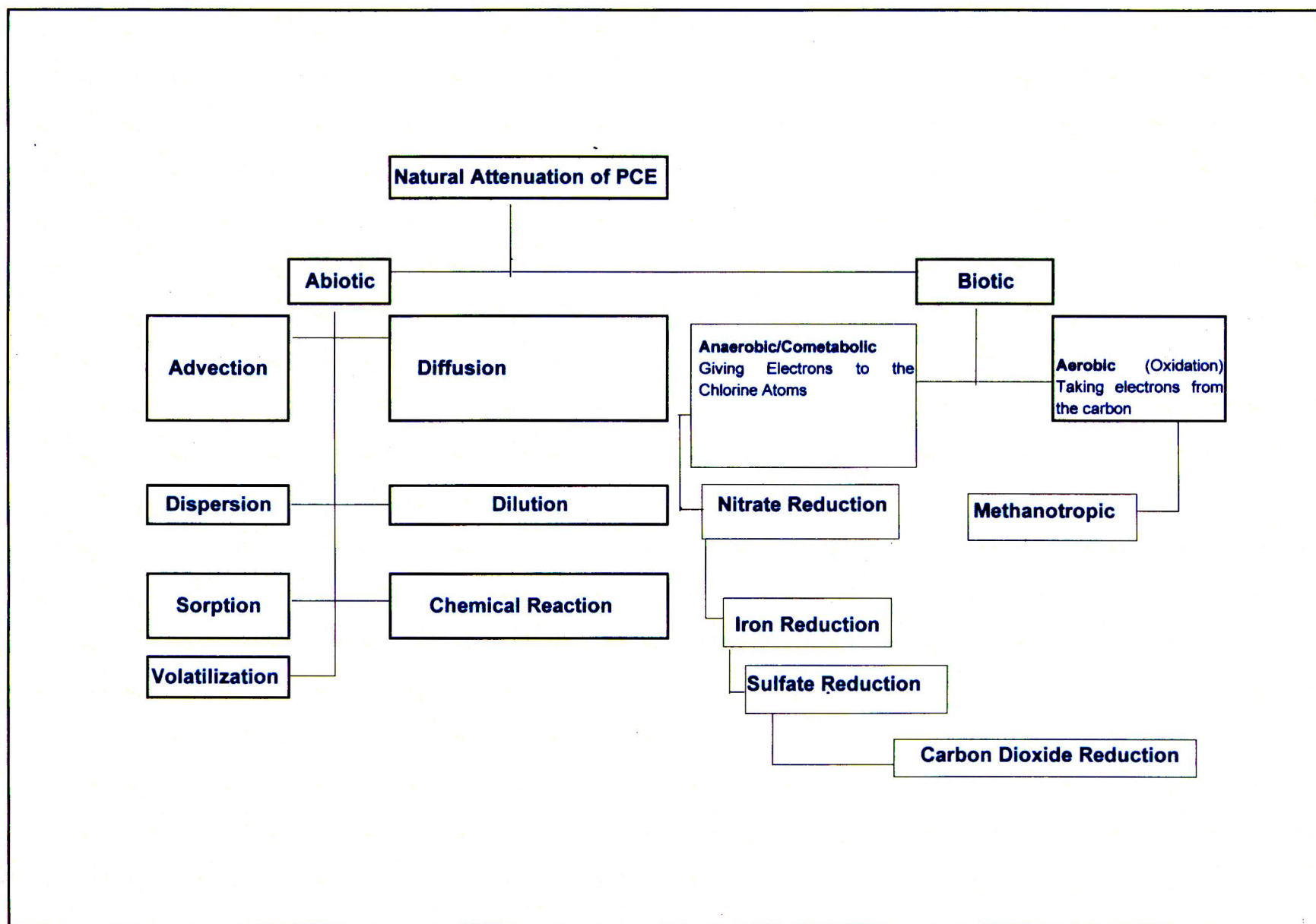
The intent of this document is to provide a resource for collection of site characterization parameters, analytical methods, sampling techniques and data evaluation procedures which may be necessary in considering natural attenuation as a remediation tool at sites contaminated with chlorinated solvents. This document does not include a detailed discussion of the natural attenuation processes; however it does present a logical approach to the evaluation, substantiation, feasibility, and appropriateness of implementing natural attenuation at a site as part of, or as the sole remedial approach. Since information from field studies and research regarding natural attenuation of chlorinated solvents is ongoing, this guide will be revised as necessary.

The applicability of implementing Monitored Natural Attenuation (MNA) as a remedial strategy at a drycleaning site will include considerations regarding the presence of free product, contaminated soil, groundwater contaminant concentrations, physical, chemical, and biological characteristics of each contaminant, trend analysis, temporary points of compliance, and overall technical evaluation. The proposed drycleaning rule, F.A.C. Chapter 62-782 provides information regarding these specific site considerations.

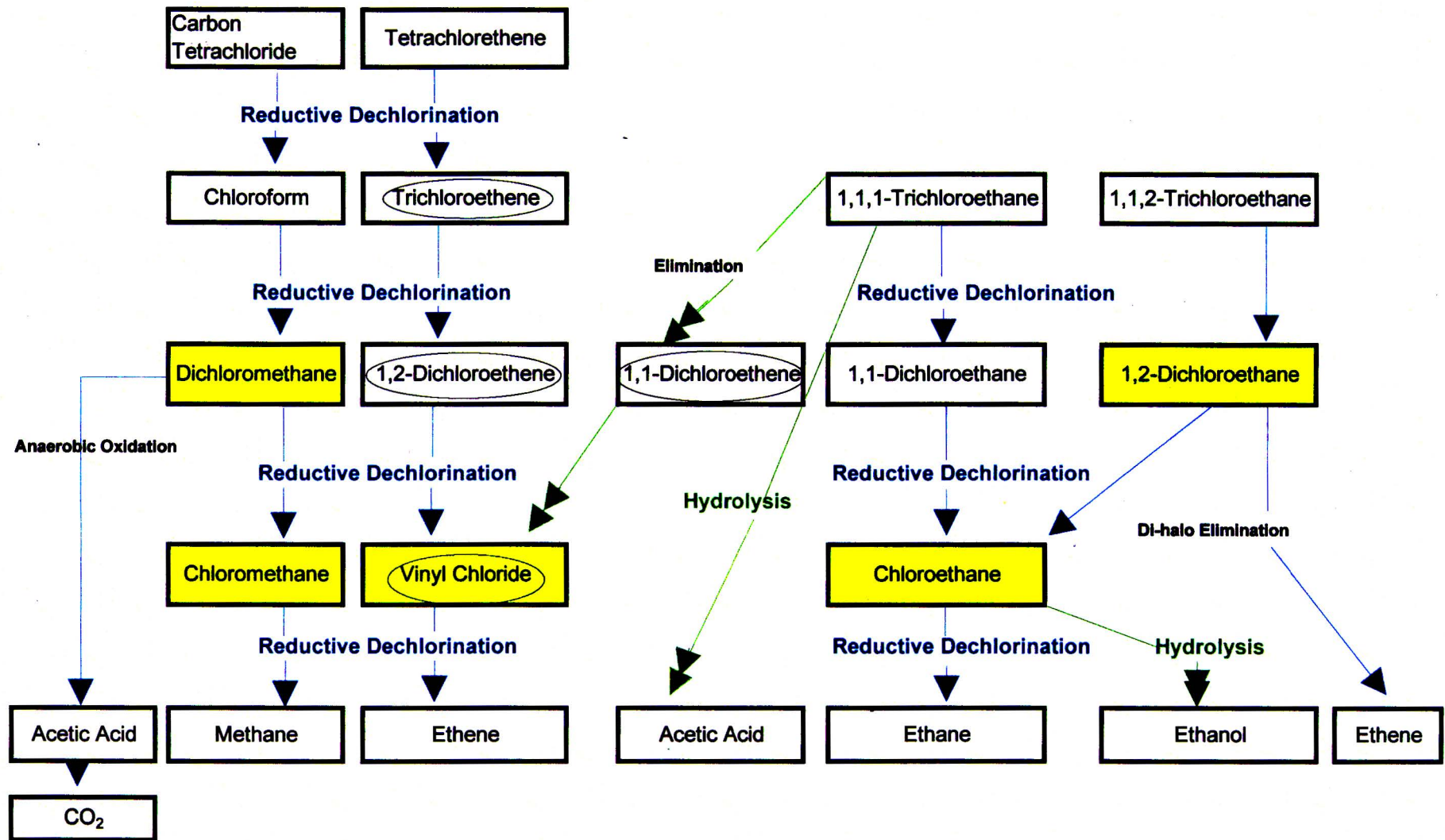
Natural attenuation refers to naturally-occurring processes in soil and groundwater that act without human intervention to reduce contamination. The processes that may contribute to plume attenuation are physical, chemical and biological. Natural attenuation is the combined effect of dispersion, dilution, volatilization, sorption, chemical reactions with subsurface materials, and biodegradation of dissolved contaminants in the groundwater and soil. Intrinsic bioremediation or biodegradation is used to describe the natural biological process that leads to contaminant degradation and is the only process that actually results in significant contaminant destruction during natural attenuation.

MNA is considered an approach to site rehabilitation that allows natural processes to contain the spread of contamination and reduce the concentrations of contaminants in groundwater and soil. This approach should be substantiated by site specific assessment information including data evaluation and interpretation, and long-term monitoring. MNA may be used in conjunction with other engineered remedies. Natural attenuation should not be considered a "no action" alternative to site rehabilitation, but rather a remedial approach which focuses on the verification and monitoring of natural remediation processes. Natural attenuation processes should be evaluated





**Figure 1**  
**Natural Attenuation of Chlorinated Solvents in Groundwater**



**Biotic reactions (Anaerobic Conditions)** →  
**Abiotic reactions (Anaerobic or aerobic conditions)** →  
 Aerobic cometabolism to CO<sub>2</sub> in presence of toluene ○  
 Aerobic mineralization to CO<sub>2</sub>

**FIGURE 2**  
**COMMON DEGRADATION PATHWAYS**

Reductive dechlorination generally occurs sequentially as illustrated above. Thus, reductive dechlorination of chlorinated solvents results in a proliferation of daughter products and an increase in the concentration of chloride ions. The rate of reductive dechlorination has been observed to decrease as the degree of chlorination decreases; vinyl chloride being the least susceptible to reductive dechlorination (13, 14). The transformation of DCE to VC and VC to ethene requires higher reducing conditions. Therefore, under specific site conditions, natural attenuation by biotic processes may not proceed to completion and, as a result, intermediate compounds may accumulate (13, 14).

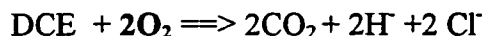
**Figure 3, Anaerobic Reductive Dechlorination**, (Cover page of document), illustrates the sequential dechlorination of a trichloroethene plume and **Figure 4, Reductive Dechlorination**, illustrates the degradation of PCE and the interactive processes of the electron donors, degradation products, and environmental conditions.

The efficiency of reductive dechlorination differs for methanogenic (CH<sub>4</sub>), sulfate-reducing (SO<sub>4</sub><sup>+2</sup>), iron-reducing (Fe<sup>+3</sup>), or nitrate-reducing (NO<sub>3</sub><sup>+2</sup>) conditions. The dechlorination of PCE and TCE to DCE will occur under milder reducing conditions. **Figure 5, Common Patterns of Chlorinated Solvent Biodegradation in an Anaerobic System**, illustrates the intricacy of the degradation process and transitional substrates and degradation products.

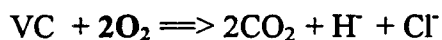
Until recently, it was commonly believed among researchers that vinyl chloride would not degrade under anaerobic conditions and would accumulate; however, it has now been documented that vinyl chloride can be degraded under both aerobic and anaerobic conditions if the necessary electron acceptors, organic carbon and subsurface microorganisms are present. Recent research has shown that there is evidence of *mineralization* of vinyl chloride under iron-reducing conditions so long as there is sufficient bioavailable ferric iron (Fe<sup>+3</sup>)(4).

### 2.3 Electron Donor Reactions - Aerobic Biodegradation

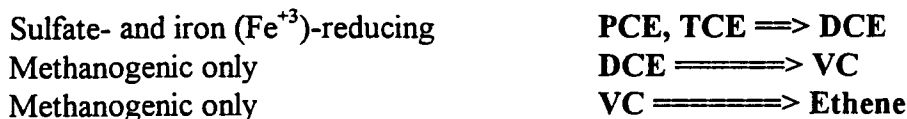
Illustrated below are two equations demonstrating that facilitating microorganisms obtain energy and organic carbon from the degraded aliphatic chlorinated hydrocarbon (20). Under aerobic conditions, DCE and VC can be oxidized to CO<sub>2</sub> (2, 19):



and



The **anaerobic** conditions include either methanogenic, sulfate-reducing or iron-reducing conditions and include the following general reactions:



The **aerobic** conditions include: **DCE, VC  $\implies$  2CO<sub>2</sub> and Cl<sup>-</sup>**



## 2.4 Co-metabolism

The degradation of a chlorinated aliphatic hydrocarbon can be catalyzed by the presence of an enzyme or co-factor that is produced by an organism for other purposes (20). For example, during co-metabolism, trichloroethene is indirectly transformed by various microbial enzymes during the biodegradation of methane, toluene, phenol, or ammonia. The organism receives no known benefit nor does trichloroethene enhance the degradation of BTEX or other carbon sources. Co-metabolism is not a stable process and has been best documented in aerobic conditions, although it can occur under anaerobic conditions (14, 15, 18, 20).

## 3.0 NATURAL ATTENUATION BY PHYSICAL PROCESSES

The primary physical processes of natural attenuation include dispersion and dilution. Field observations have indicated that contaminant concentrations decline in a pattern which is consistent with the soil and/or groundwater processes. For example, groundwater flow direction and high hydraulic conductivity would indicate contaminant concentration decline along a particular flow path. It has been documented that both dispersion and dilution processes exhibit minor depletion of oxygen and redox potential values which may occur in the source area. However, values for most other field and inorganic parameters will not change significantly. Generally, site characteristics which may facilitate the attenuation processes include coarse sand and gravel, high groundwater flow and shallow fractured bedrock systems (28).

## 4.0 METABOLIC BY-PRODUCTS AND REDOX PROCESSES IN GROUNDWATER SYSTEMS

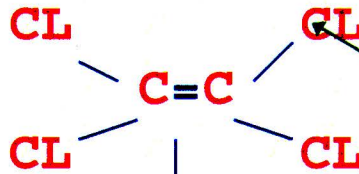
Redox potential (ORP) is a measure of electron activity and an indicator of the tendency of a solution to accept or transfer electrons. Groundwater reactions involving chlorinated solvents are usually biologically mediated, and the oxidation-reduction potential depends on and influences the rate of biodegradation. Additionally, some biological processes operate only within a specific range of redox conditions. The possible redox potential for reductive dechlorination ranges from approximately -400 to 800 millivolts (mV). **Figure 6, Redox Potential**, indicates typical redox conditions for groundwater with different electron acceptors.

It is evident that one of the keys in understanding and evaluating the biodegradation of chlorinated ethenes is to provide accurate delineation of redox conditions and metabolic by-product concentrations in the groundwater at the site. Listed below are several generalized equations which indicate several common redox reactions in groundwater: (2)

<u>REACTION</u>	<u>SYSTEM TYPE</u>
1. $O_2 + CH_2O \Rightarrow CO_2 + H_2O$ • Solubility of oxygen in water is ~10 mg/l	<b>Aerobic Respiration</b> <i>Measured by <math>O_2</math> depletion and <math>CO_2</math> production.</i>

# REDUCTIVE DECHLORINATION

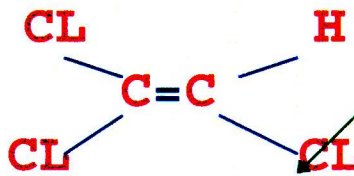
PCE



H<sup>+</sup>(2e)

ELECTRON DONOR

TCE

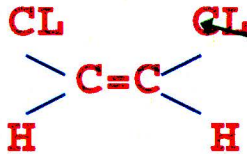


H<sup>+</sup>(2e)

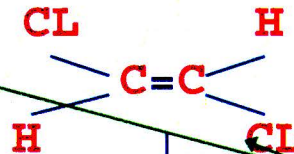
ELECTRON DONOR

cis 1,-2 DCE

BIOLOGICAL DCE IS  
80-100%

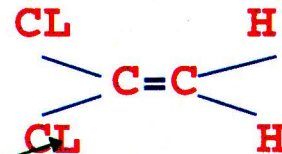


trans 1,2 DCE



1,1 DCE

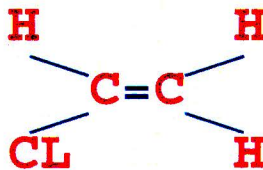
MANUFACTURED DCE  
IS MOSTLY 1,1 DCE



H<sup>+</sup>(2e)

ELECTRON DONOR

VINYL CHLORIDE



AEROBIC ENVIRONMENT  
MUST HAVE ELECTRON  
ACCEPTORS:  
OXYGEN  
NITRATE

ANAEROBIC  
ENVIRONMENT

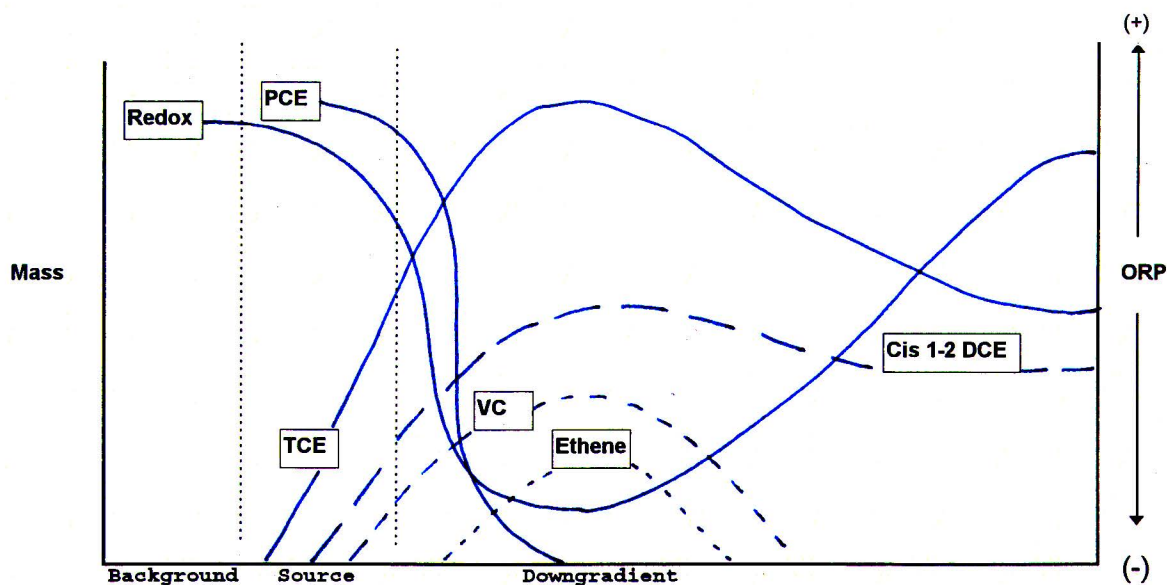
EASIEST- RATE OF REACTION  
CONTROLLED BY CARBON  
SOURCE

HARDEST

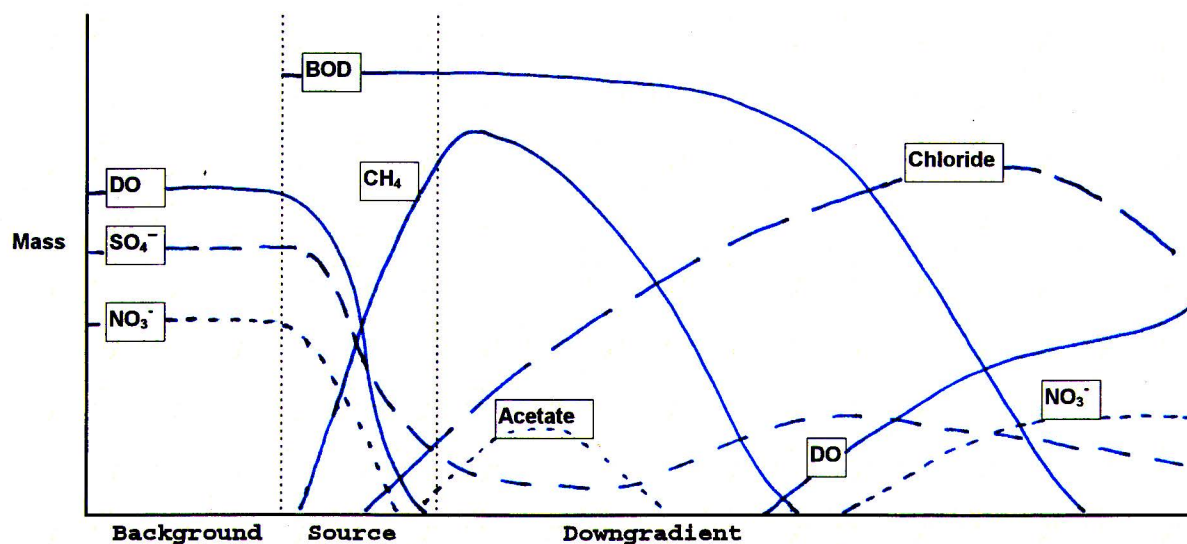
AEROBIC  
ENVIRONMENT

NOTE: VC MAY BE MINERALIZED  
UNDER ANAEROBIC CONDITIONS

REDUCTIVE DECHLORINATION  
FIGURE 4

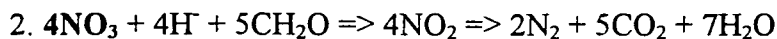


In the anaerobic zone PCE is dechlorinated to TCE, DCE, VC and Ethene. The dechlorination rate is insufficient to cause all of the TCE and DCE to be dechlorinated in the anaerobic zone. These chemicals along with methane, ethene and vinyl chloride migrate into the transition and aerobic zone.



Due to the metabolism in the source area, an anaerobic zone has developed in the groundwater system. Methanogenic, sulphate-reducing, and acetogenic bacteria are active to cause TCE and DCE to be dechlorinated in the anaerobic zone. Several geochemical parameters are good indicators of bacteria metabolism. Ethene and VC are mineralized to CO<sub>2</sub> by aerobic bacteria in the aerobic zone. (Adapted from ITRC-Natural Attenuation of Chlorinated Solvents in Groundwater-Training Course Workbook.)

**FIGURE 5**  
**COMMON PATTERNS OF CHLORINATED SOLVENT BIODEGRADATION IN A SEQUENTIAL AEROBIC/ANAEROBIC SYSTEM (JK)**



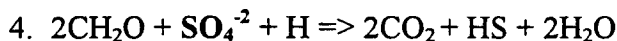
- Solubility of nitrate almost unlimited in water
- Natural groundwater range <1 to 10 mg/l
- Groundwater  $\text{NO}_3$  concentration is usually higher due to anthropogenic inputs such as fertilizer.

**Nitrate Reduction**  
*Measured by  $\text{NO}_3$  depletion and  $\text{NO}_2$  production.*



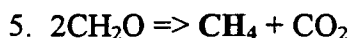
- Solubility of  $\text{Fe}^{+3}$  (ferric iron) is low; normal measured groundwater  $\text{Fe}^{+3}$  concentrations are <1 mg/l. However, an almost unlimited supply may be available in the aquifer material.

**Iron Reduction**  
*Measured by  $\text{Fe}^{+3}$  depletion and  $\text{Fe}^{+2}$  production.*



- Natural sulfate concentrations are usually >5 mg/l. Sulfate may be higher due to human inputs.

**Sulfate Reduction**  
*Measured by  $\text{SO}_4^{-2}$  depletion and HS production.*



- Supply of  $\text{CO}_2$  is unlimited. However, only specific electron donors can supply energy for this reaction and specific conditions are required.
- The solubility limit of  $\text{CH}_4$  in water is ~40 mg/l.

**Methanogenesis**  
*Measured by  $\text{CH}_4$  production*

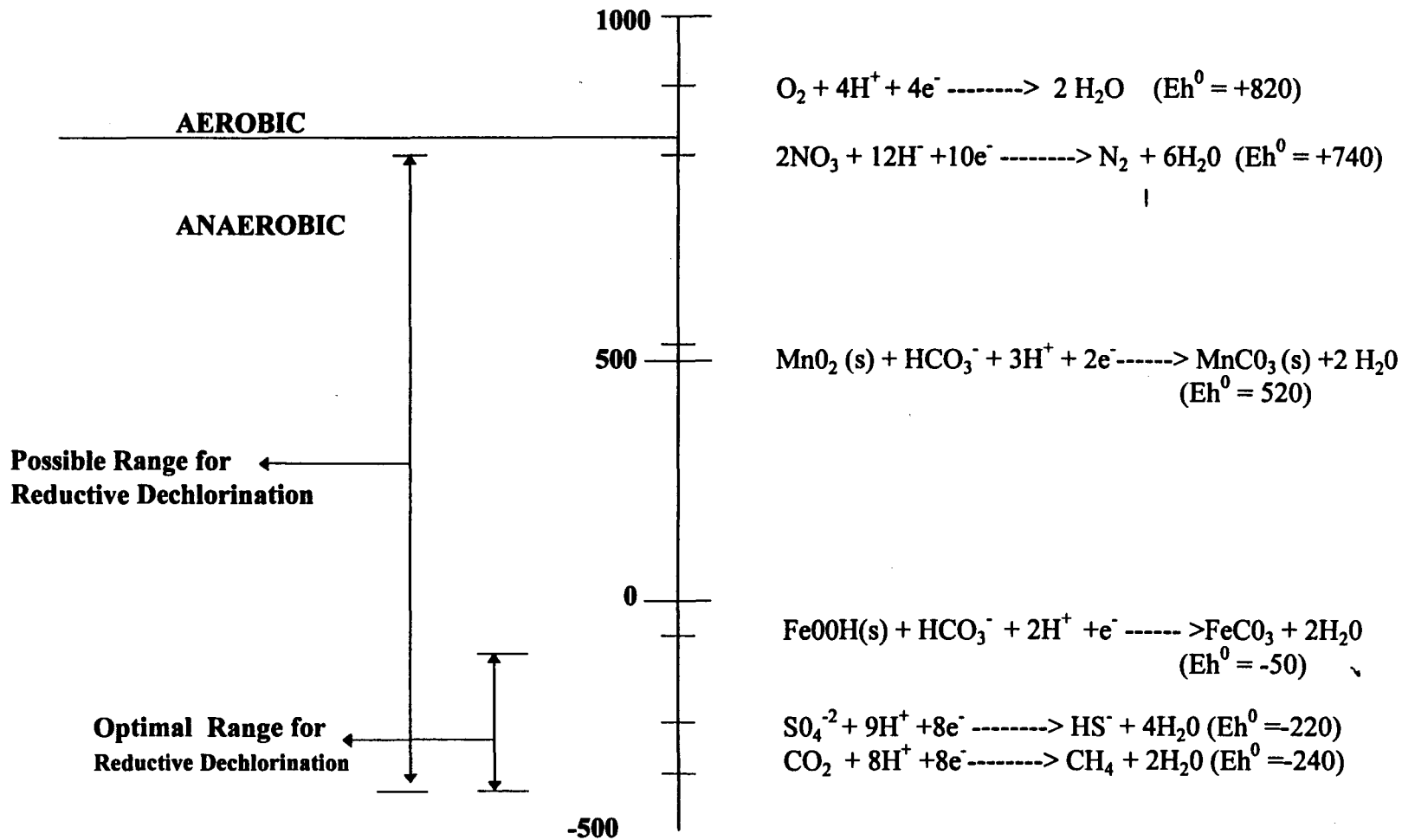
The methodology for assessing redox conditions and metabolic by-products involves tracking the **disappearance** of electron acceptors and the **appearance** of end products. For example, nitrate and sulfate concentrations below background levels in the plume would indicate anaerobic biodegradation through denitrification and sulfate reduction. Elevated concentrations of metabolic by-products such as ferrous iron and methane may indicate the occurrence of ferric iron reduction and methanogenesis inside the plume.

Oxidation-reduction results can be used to contribute real-time data on the location of the contaminant plume, and mapping these areas will help to determine the approximate location of the plume.

## 5.0 EVIDENCE USED TO SUPPORT NATURAL ATTENUATION

MNA may be proposed as a cleanup strategy at sites that meet certain criteria as set forth in the proposed drycleaning rule, F.A.C. Chapter 62-782. Depending on site specific conditions, including contaminant concentration levels, and plume location, evidence may be required to provide a projection of the potential decrease in the extent of the contamination plume and furnish sufficient evidence to demonstrate what natural attenuation mechanisms are occurring at the site. There is a range of levels of effort associated with the demonstration of monitored natural attenuation as a cleanup strategy. An incremental approach may include the following lines of evidence:

**Redox Potential (Eh<sup>0</sup>)  
in Millivolts @ pH=7  
and T=25<sup>0</sup>C**



**FIGURE 6- REDOX POTENTIAL FOR VARIOUS ELECTRON ACCEPTORS**  
(Modified from Bouwer 1994-Draft EPA Region 4 "Suggested Practices for Evaluation of a Site for Natural Attenuation..." 1997)

### **1. Direct evidence of contaminant concentration reduction and degradation products**

This is simply an observed reduction in the concentration of released contaminants downgradient from the source area over time. However, biological processes may be indicated by an increase in the concentration of some degradation products, such as vinyl chloride. **Figure 5**, illustrates degradation products generated through biotic processes as related to distance and source area.

Typically, this first line of evidence may be documented by reviewing historical trends in contaminant concentrations and distribution in conjunction with site geology and hydrogeology to show reduction in the total mass is occurring at the site. Other site specific conditions, including hydraulic properties and aquifer characteristics determined from the contamination assessment, will also provide valuable information regarding groundwater flow and transport. Evaluation of all data should establish if the plume is expanding, stable or shrinking.

### **2. Examine changes in the concentration and distribution of the geochemical and biochemical indicator parameters that are related to specific natural attenuation processes**

This information, along with the migration and contaminant degradation products, provide specific data to support natural attenuation. This line of evidence relies on chemical and physical data to show that contaminant mass is being destroyed via biodegradation and/or attenuated by physical processes such as diffusion, dilution, sorption, dispersion and other chemical reactions. The collection and analysis of the geochemical parameters is site specific and should be evaluated carefully. All geochemical parameters may not be necessary at each site.

### **3. Laboratory microcosm studies**

These studies provide data which may be used to estimate site specific biodegradation rates which cannot be demonstrated by field data alone and also to confirm specific chlorinated solvent biodegradation processes. Evaluation of research data has indicated that this line of evidence is generally not cost effective.

## **6.0 METHODOLOGY USED FOR DATA REVIEW AND SITE EVALUATION**

Evaluation of existing data should be used as an initial indicator of natural attenuation (e.g., evidence of transformation, appearance of daughter products and mass loss of target chemicals). Where there is initial evidence of natural attenuation mechanisms, further evaluation with regard to the type, rate and effectiveness of these mechanisms may be accomplished by the collection and analysis of additional specific natural attenuation parameters. Completion of the following steps will help facilitate a logical and sequential quantification process throughout the entire evaluation process:

#### **Step # 1 Site Characterization Data**

Review all site characterization data which will be used to develop a site specific conceptual model. For the evaluation of biotic attenuation, collection of specific parameters may be required to determine which microbial processes predominate in the site specific groundwater system.

Major geochemical parameters that characterize the subsurface include: dissolved oxygen, alkalinity, pH, redox potential, temperature, and concentration of electron receptors. **Tables 1 and 2** provide a list and interpretation of several parameters.

**TABLE 1 - Geochemical Parameters (7)**

<u>Parameter</u>	<u>Range</u>	<u>Interpretation</u>
Redox potential	<50 mV	Reductive pathway possible
Sulfate	<20 mg/L	Competes at higher concentrations with reductive pathway
Nitrate	<1 mg/L	Competes at higher concentrations with reductive pathway
Oxygen	<0.5 mg/L	Tolerated; toxic to reductive pathway at higher concentrations
Oxygen	>1 mg/L	Vinyl chloride oxidized
Iron (II)	>1 mg/L	Reductive pathway possible
Sulfide	>1 mg/L	Reductive pathway possible
Hydrogen	>1 nM	Reductive pathway possible; vinyl chloride may accumulate
pH	5<pH<9	Suitable for biological activity

Evaluation of the geochemical parameters and other site information will contribute to understanding the biotic natural attenuation processes. Listed below are some **Geochemistry Rules of Thumb (27,28)** which can also be used to help evaluate data quality:

- Dissolved oxygen (DO) concentrations are always below 10 ppm.
- DO levels are directly proportional to redox potential. (Relative)
- DO is inversely proportional to Fe<sup>+2</sup> and alkalinity concentrations.
- Alkalinity concentrations are directly proportional to Fe<sup>+2</sup>, but Fe<sup>+2</sup> concentrations are not necessarily direct to alkalinity.
- Methane is always below 25 ppm.
- Dissolved iron concentrations over 5 ppm may establish the reducing pathway.
- Coastal aquifers are normally oxidized.
- Biological DCE is at least 80% cis-1,2 DCE.

**TABLE 2 - Contaminants and Daughter Products (7)**

<u>Parameter</u>	<u>Interpretation</u>
PCE	Material spilled
TCE	Material spilled or daughter product of perchlorethylene (PCE)
1,1,1-Trichloroethane	Material spilled
<i>cis</i> -DCE	Daughter product of trichloroethylene (TCE)
<i>trans</i> -DCE	Daughter product of trichloroethylene (TCE)
Vinyl Chloride	Daughter product of dichloroethylene (DCE)
Ethene	Daughter product of vinyl chloride
Ethane	Daughter product of ethene or chloroethane
Methane	Product of methanogenesis or reductive dechlorination of chlorinated methanes, such as carbon tetrachloride, chloroform and methylene chloride
Chloride	Daughter product of organic chlorine
Carbon dioxide	Ultimate oxidative daughter product
Alkalinity	Results from interaction of carbon dioxide with aquifer minerals and other factors such as the biomass using the inorganic carbon in HCO <sup>3-</sup> as a carbon source.

Several **Rules of Thumb for Contaminant Half Lives** (28), will assist in understanding the time frames and processes that are occurring within the plume boundary.

- BTEX contaminants have a half life that is very short ( weeks).
- Chlorinated Hydrocarbon's half life are months to years.
  - ◊ DCE is the slowest to biodegrade. It's half life will determine if natural attenuation is biologically mediated.
  - ◊ VC can degrade very fast under the proper conditions.

The September 1998 *Considerations for Assessment of Drycleaning Solvent Contaminated Sites* (21), provides general guidance regarding Contamination Assessment (CA) activities for sites under the Drycleaning Solvent Cleanup Program. The goal of any site characterization is to define the extent of contamination and to better understand the fate and transport of the contaminants. This is necessary in order to assess any current or potential threat to human health and the environment. The data elements in **Table 3** provide a review of the key information necessary for plume delineation. This information may also be necessary to substantiate the effectiveness of natural attenuation through both physical and biological processes.

**TABLE 3 - Data Elements**

<u>Data</u>	<u>Information</u>
Direction and gradient of groundwater flow	Estimate expected rate and direction of plume migration
Hydraulic Conductivity	Estimate expected rate of plume migration
Aquifer thickness	Model groundwater flow
Analysis of site stratigraphy	Conceptual model for preferential flow paths for contaminant transport
Water table fluctuations	Potential source smearing and variation in flow direction
Date of contaminant release	Estimate anticipated extent of plume migration
Contaminant delineation- soil and groundwater	Current status of plume
Soil texture, structure, TOC content	Higher organic carbon content and smaller grain size results in greater adsorption of chemicals and retardation of migration
Historical concentrations along flow paths	Plume status (i.e., migrating, decreasing, steady state)
Locations of groundwater recharge areas	Identify areas of ground water aeration
Locations of exposure receptors	Identify possible exposures

**Step #2 Three-dimensional Conceptual Model**

The existing site data gathered in the site characterization phase may be used to develop a three-dimensional (3D) **conceptual model** (27). Both physical and biological components of natural attenuation are incorporated into the conceptual model. The model should therefore include a



representation of the groundwater flow and transport system based upon the geological, biological, geochemical, hydrological, climatological, and analytical data. Use of this model will represent the site conditions and help to identify data shortcomings. If required, the data collected during site characterization may also be used to simulate the fate and transport of contaminants. This simulation will help to illustrate predictions regarding the future extent and concentrations of contaminants in the dissolved plume.

### **Step #3 Additional Site Characterization Data Collection**

Examine the site data for evidence of natural attenuation and develop an hypothesis (conceptual model) to explain the attenuation process. With use of mobile laboratories and direct push technology, the initial conceptual models can change dramatically while in the field. It is imperative that data are carefully reviewed and evaluated to constantly reconstruct the conceptual model. Select locations for additional data collection and evaluate what specific data are necessary to complete the conceptual model. Some common elements are: chlorinated solvent distribution along flow pathway, presence of degradation products, evidence of mass loss of chemicals, and evidence of geochemical or biochemical indicators of natural attenuation.

During site characterization, identify any source areas that are suspected. Source removal usually is very effective at decreasing the remediation time frame.

### **Step #4 Identify Exposure Pathways**

### **Step #5 Refine the Site Conceptual Model**

Perhaps one of the most important steps in the process is to incorporate all new data into the site conceptual model. Once the conceptual model is complete, a remedial alternative analysis may be initiated for site rehabilitation.

## **7.0 NATURAL ATTENUATION SAMPLING AND DATA EVALUATION METHODOLOGY**

The main objective of the natural attenuation evaluation is to determine whether natural attenuation processes are taking place, and if those processes will reduce contaminant concentrations in groundwater within an acceptable distance and time frame. Upon examination of the initial site contaminant data in the field, if degradation products exist (i.e., daughter products), it is reasonable to infer that there is biotic natural attenuation. If historical data is available, reduction of contaminant concentrations in groundwater, without degradation products, may indicate abiotic natural attenuation processes. The primary natural attenuation mechanism may now be identified and substantiated through sampling, analysis, and evaluation of parameters that relate to each significant natural attenuation mechanism.

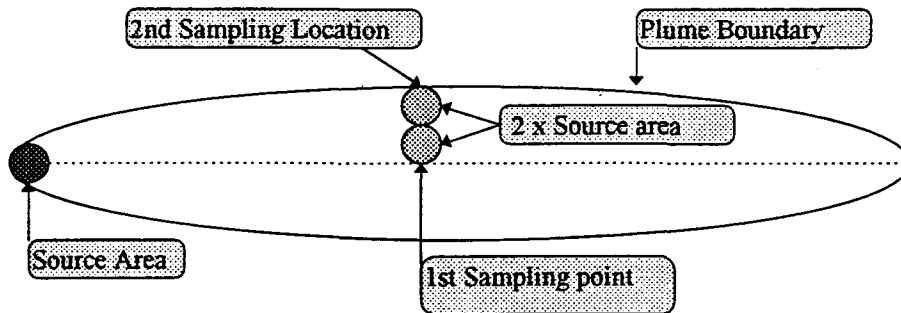
The following may be used to contribute toward increased assessment efficiency:

- Utilization of direct push technology, whenever possible, will assist in mapping the plume both laterally and vertically.

- Field test kits may be used for the following parameters: DO (dissolved oxygen), Fe<sup>+2</sup>(ferrous iron), alkalinity, sulfide, pH, and temperature.
- Mobile laboratories will streamline site evaluation with real time information.

The use of temporary transects for initial site natural attenuation characterization (27) may be helpful in identifying plume geochemical conditions. The placement of these transects may be useful for identifying **both biotic and abiotic natural attenuation processes**. Indicated below is an example of an elongated groundwater plume with the placement of sampling transects located in the proper locations.

- ◇ Locations may be determined by starting at the inferred center of the plume and moving out in a stepwise fashion at intervals of 2 times the source area (inferred) width.
- ◇ If the 2nd sampling location is contaminated, then sample 2 times the source area width further along the transect.



### Vertical Profiling

- The goal is to determine the variations in the physical and biological systems throughout the site (27, 28).
- One of the most important physical characteristics to determine at a site is **hydraulic conductivity** (27, 28).

Identify the significant biodegradation processes in source, plume, fringe, and background area from review of all collected data.

### 7.1 Monitor Well Placement and Sampling for MNA

Monitor well placement will depend on specific site conditions and may be installed upon completion of all assessment activity. The following are typical locations, but may be adjusted depending on the site conditions:

1. **In the most contaminated zone.** This should be as close to the source area as possible to provide a determination of the dominant terminal electron-accepting processes occurring at the site.

2. **Downgradient from the source area and within the dissolved plume.** This will provide information as to whether the plume is degrading with distance along the migration path and identify the key geochemistry components. This information will also provide contaminant flux estimates along the plume pathway. The number of monitor wells located along the contaminant pathway may be determined by the site specific plume length.
3. **Downgradient of the dissolved plume.** This will provide unaffected groundwater analysis information.
4. **Upgradient and lateral locations not affected by the plume.** This will provide additional background concentrations of specific geochemistry indicators and laterally define plume boundary.

If required, soil parameters may also be taken as noted in **Table 4, Natural Attenuation Parameters and Test Methods**, to provide data necessary for solute transport modeling.

Groundwater samples at **selected monitor locations** may be analyzed for some or all of the parameters in **Table 4**. The table provides **guidance** regarding specific parameters and methods that may be included during the initial sampling and assessment period. Upon subsequent evaluation of the data, several parameters may be removed from future sampling events.

If several aquifers have been affected by the contaminant plume, and if degradation products are present, the sampling plans should be flexible to include the multiple plumes and zones of contamination. Sampling for specific parameters should always integrate overall cost effectiveness and appropriate site management.

## **7.2 Evaluation Of Biological Mediated Natural Attenuation Data**

Strength of evidence for reductive dechlorination is based upon the evaluation of the contaminants and the parameters as listed in **Table 5, Analytical Parameters and Data Evaluation**. The information in **Table 5** relies on the fact that biodegradation will cause predictable changes in groundwater chemistry. The evaluation of all data will provide information to determine if, and what type of, **biologically mediated** natural attenuation is taking place at the site.

## **8.0 MONITORING THE EFFECTIVENESS OF NATURAL ATTENUATION**

The conceptual model should now include both physical and biological natural attenuation processes. Periodic monitoring will be necessary to determine whether the predictions of the site characterization are accurate and to ensure that all receptors are protected.

### **8.1 Estimation Of Biodegradation Rate**

Depending on site complexity, and plume location, a natural attenuation rate model may be incorporated into the *MNA* process. Biodegradation of chlorinated aliphatic hydrocarbons have been represented as a first order decay, with the rate being represented by the half-life:

TABLE 4

MATRIX	PARAMETER	TEST METHOD	LOCATION	COMMENTS	JUSTIFICATION
Soil	Volatile Organics	EPA 8021 with updated EPA 5035 preparation	Fixed	Must determine high and low levels.	Determine the extent of soil contamination.
Soil	Total Organic Carbon (TOC)	SW-846 9060 modified	Fixed	Procedure must be accurate over the range of 0.1% to 5% TOC.	The amount of TOC in aquifer matrix influences contaminant migration and/ or degradation.
Soil gas	Volatile Organics	EPA T0-14	Fixed	****Not recommended at all sites.	Distribution of soil gas within plume area.
Water	Alkalinity	HACH / EPA 310.1/310.2	Field / Fixed	Collect 100 to 250 ml of water in a glass container.	Provides an indication of the buffering capacity of groundwater and the amount of carbon dioxide dissolved; increases due to biodegradation of organic compounds.
Water	Chloride	HACH / EPA E-300 series	Field / Fixed	Colorimetric Field Kit	Final product of chlorinated solvent reduction.
Water	Dissolved Oxygen	Meter and Probe Flow through cell 360.1  Winkler titration ASTM D 888-92 (A)	Field	With all DO methods extra care must be taken to avoid aeration during all steps of the analysis including well purging.  Field kits for performing Winkler titrations can be used as the primary method of DO measurement or to confirm meter measurements.	Concentration <1mg/L = anaerobic pathway; highest energy-yielding electron acceptor for biodegradation of organic constituents.
Water	Ethane, Ethene, Methane	SW-846 8015 modified	Fixed	Method published by EPA researchers. Limited to few commercial labs.	Presence of methane suggests biodegradation of organic carbon via methanogenesis; ethane and ethene are daughter products of chlorinated solvents. ***Not recommended for all sites.
Water	Ferrous Iron (Fe <sup>+2</sup> )	HACH (25140-25) / EPA E-300 series	Field / Fixed	Colorimetric Field Kit Collect 100 mL of water in glass container. Filter sample with 0.2 micron filter if turbid.	Reductive pathway; anaerobic degradation process due to depletion of oxygen, nitrate and manganese; increased concentrations of Fe <sup>+2</sup> may indicate Fe <sup>+3</sup> is being used as an electron acceptor during anaerobic biodegradation.
Water	Hydrogen	Not recommended at this time.			Under consideration.

TABLE 4  
NATURAL ATTENUATION PARAMETERS AND TEST METHODS

TABLE 4

Water	Redox Potential	Field Flow through cell	Field	ORP probe can be inserted into flow through cell and reading obtained simultaneously with DO, pH, T, and EC.	Indicates the relative ability of a solution to accept or transfer electrons; define regions of the plume under oxidizing and reducing conditions; helps validate DO measurements. May range from more than 800mV for aerobic conditions to -400mV.
Water	Temperature	Field Probe / EPA 170.1	Field / Fixed	Available from some DO, ORP, pH, or EC probes.	At T>20, biochemical process accelerated; proper well development index.
Water	Nitrate	HACH / EPA E-300 series	Field / Fixed	Collect 40 mL of water in a glass or plastic container; add sulfuric acid to pH<2, keep cool.	Substrate for microbial respiration if oxygen depleted; decreased nitrate concentrations in anaerobic portion of the plume may indicate use of sulfate as an electron acceptor.
Water	pH	EPA 150.1 or SW-846 9040	Field	Direct Reading Meter	Difference in pH between contaminated and uncontaminated groundwater may indicate biological activity is occurring.
Water	Sulfate (SO <sub>4</sub> ) <sub>2</sub>	HACH / EPA 8051	Field / Fixed	Colorimetric Method	Substrate for anaerobic microbial respiration; decreased sulfate concentrations in anaerobic portion of the plume may indicate use of sulfate as an electron acceptor.
Water	Sulfide	HACH / EPA E-300 series	Field / Fixed	Colorimetric Method	Microbially reduced form of sulfate. Indicates reduced conditions.
Water	Total Organic Carbon (TOC)	HACH / EPA E-300 series	Field / Fixed		A measure of the total concentration of organic material in water that may be available for biological degradation.
Water	Volatile Organics	EPA 8021 / 8015	Fixed / Mobile	Choice of methods --cost effective considerations required.	Provides a measure of the type and quantity of parent and biogenic daughter products.

**Notes:** SW-846 refers to the *Test Methods for Evaluating Solid Waste, Physical and Chemical Methods*  
 EPA refers to *Methods for Chemical Analysis of Water and Wastes*  
 HACH refers to the Hach Company catalog

TABLE 4  
 NATURAL ATTENUATION PARAMETERS AND TEST METHODS

Analyte	Concentration	Interpretation
Oxygen	<0.5 mg/L	Tolerated: suppresses reductive dechlorination at higher concentrations
Oxygen	>1 mg/L	Vinyl chloride may be oxidized aerobically, but reductive dechlorination will not occur
Nitrate	<1 mg/L	At higher concentrations may compete with reductive pathway
Iron (II)	>1 mg/L	Reductive pathway possible; vinyl chloride may be oxidized under Iron(III)-reducing conditions
Sulfate	<20 mg/L	At higher concentrations may compete with reductive pathway
Sulfide	>1 mg/L	Reductive pathway possible
Methane	<0.5 mg/L	Vinyl chloride oxidizes
	>.5 mg/L	Ultimate reductive daughter product, vinyl chloride accumulates
Oxidation-Reduction	<50mV	Reductive pathway possible
	<-100mV	Reductive pathway likely
pH	5<pH<9	Optimal range for reductive pathway
	5>pH>9	Outside optimal range for reductive pathway
TOC (Total Organic Carbon)	>20 mg/L	Carbon and energy source; drives dechlorination; can be natural or anthropogenic
Temperature	> 20°C	At T>20°, biochemical process is accelerated
Carbon dioxide	> 2 X background	Ultimate oxidative daughter product
Alkalinity	> 2 X background	Results from interaction of carbon dioxide with aquifer minerals
Chloride	> 2 X background	Daughter product of organic chlorine
Hydrogen	>1nM	Reductive pathway possible; vinyl chloride may accumulate
	<1 nM	Vinyl chloride oxidized
Volatile Fatty Acids	>0.1 mg/L	Intermediates resulting from biodegradation of aromatic compounds; carbon and energy source
BTEX	>0.1 mg/L	Carbon and energy source; drives dechlorination
Perchloroethene		Material released
Trichloroethene		Material released or daughter product of perchloroethene
Dichloroethene		Material released or daughter product of trichloroethene If amount of cis-1,2 DCE is greater than 80% of total DCE, it is likely a daughter product of trichloroethene. 1,1-DCE can be chemical reaction product of TCA.
Vinyl Chloride		Daughter product of dichloroethene
Ethene	>0.01 mg/L	Daughter product of vinyl chloride
Ethane	>0.1 mg/L	Daughter product of ethene
Chloroethane		Daughter product of vinyl chloride under reducing conditions
Dichloroethane		Daughter product trichloroethene

**Table 5**

**Analytical Parameters and Data Evaluation**

Adapted from "Draft EPA Region 4 Suggested Practices for Evaluation of a Site For Natural Attenuation..." 1997

$$C_0 = C_i e^{-kt}$$

Where: \*C<sub>i</sub> = initial measured chlorinated compound concentration (mg/L)  
 \*C<sub>0</sub> = final measured chlorinated compound concentration (mg/L)  
 \*t = time between initial and final measurements (year)  
 \*k = degradation rate (per year)

Two methods for determining rate constants are described by Wiedemeier et al (17).

## 8.2 Estimation of Overall Natural Attenuation Rates

Estimation of site specific attenuation rate should include all loss mechanisms that contribute to concentration changes in target contaminants, such as dispersion, dilution, sorption, volatilization, **and** biodegradation. The most compelling substantiation of a contaminant's half-life is through the evaluation of historical data. Since this may not be available at many sites, historical data may be generated by quarterly sampling for a period of one year. If no mathematically significant change in concentrations is observed, the half-life is likely very large and natural attenuation processes may not be defensible.

A method used by Buscheck and Alcantar (25) for natural attenuation rate uses interpretation of a steady-state analytical solution to the advection-dispersion equation by Bear (26). Natural attenuation rates should be calculated along the flow vector of the plume in order to show the natural attenuation process along the flow path.

For sites with complex site conditions (multiple sources, complex hydrogeology, preferential pathways, etc.), a scientific evaluation consisting of a fate and transport model may be appropriate for proposed MNA. A screening model, such as BIOSCREEN, can simulate biological degradation of hydrocarbons and has been suggested for use at chlorinated solvent sites. The model is non-proprietary and available from Robert S. Kerr Laboratory's home page ([www.Epa.gov/ada/kerriab.htm](http://www.Epa.gov/ada/kerriab.htm)). Screening models incorporating biodegradation kinetics for natural attenuation of chlorinated solvents are currently under development.

## 9.0 MONITORED NATURAL ATTENUATION REPORTS

If there is no historical groundwater monitoring data information, quarterly groundwater sampling may be required for the first year and annual evaluation of the data will determine future sampling frequency and sampling parameters. All reports should include a discussion regarding the continued applicability of MNA for the particular site. The following items may be required for inclusion in quarterly or semi-annual reports:

1. Water Table Evaluation Data
2. Groundwater Quality Data  
Subsequent review and evaluation of the data may indicate that certain sampling parameters may not be necessary for future sampling events.

3. Site maps indicating natural attenuation parameters, contaminant concentrations, specific site conditions, and neighboring areas.
4. Groundwater Historical Analytical Summary Table and graphs depicting trend analysis. The tables should include all historical information.
5. Evaluation and discussion of the data to include:
  - Geochemical evaluation
  - Contaminant trend evaluation

**Annual Reports may also include:**

*Chemical Analysis Evaluation*

- Biodegradation analysis and regression analysis may be required depending on site specific data and plume status;
- Relative mass of contaminants-plume dynamics;
- Contaminant flux concentrations along the plume;
- Review of all historical site data contaminants.

*Risk Analysis Evaluation*

- Review of property records to ascertain property owner status of the affected property;
- Review of the “physical status” of the site, and other surrounding property conditions. “Physical status” will include building development, installation of groundwater wells, residential development or other conditions that would change the exposure pathways.



## REFERENCES

1. Chapelle, F.H. U.S. Geological Survey. Protocol for Assessing the Natural Attenuation of Chlorinated Ethenes in Groundwater Systems. Draft Document July, 1996
2. Chapelle, F.H. U.S. Geological Survey. Identify Redox Conditions that Favor Natural Attenuation of Chlorinated Ethenes in Contaminated Ground Water. Symposium on Natural Attenuation of Chlorinated Organics in Groundwater. 1996 EPA/540/R-96/509
3. Vogel, T.M., C.S. Criddle, and P.L. McCarty, 1987. Transformations of Halogenated Aliphatic Compounds. *Environ. Sci. Technology* 21:722-736
4. Bradely, P.M., and F.H. Chapelle, 1996. Anaerobic Mineralization of Vinyl Chloride in Fe(III)-reducing sediments. *Environ Sci. Technology* 30:2084-2086.
5. Major, D.W., W.W. Hodgins, and B.J. Butler, 1991. Field and Laboratory Evidence of In situ Biotransformation of PCE to ethene and ethane. In: Hinchee, R.E. and R.F. Olfenbuttel, Eds. *On Site Bioreclamation*. Stoneham, Ma: Butterworth-Heinemann pp. 147-171.
6. Gossett, J.M., Zinder, S.H. Microbiological Aspects Relevant to Natural Attenuation of Chlorinated Ethenes. Symposium on Natural Attenuation of Chlorinated Organics in Groundwater. 1996 EPA/540/R-96/509
7. Wislon, B.H., Wilson, J.T. U.S. EPA National Risk Management Research Laboratory. Design and Interpretation of Microcosm Studies for Chlorinated Solvents. Symposium on Natural Attenuation of Chlorinated Organics in Groundwater. 1996 EPA/540/R-96/509
8. Cherry, J.A., University of Waterloo, Department of Earth Science, Waterloo, Ontario. Conceptual Model for Chlorinated Solvent Plumes and Their Relevance to Intrinsic Remediation. Symposium on Natural Attenuation of Chlorinated Organics in Groundwater. 1996 EPA/540/R-96/509.
9. Cherry, J.A., J.F. Barker, S. Feenstra, R.W. Gillham, D.M. Mackay, and D.J.A. Smyth. *The Borden Site for Groundwater Contamination Experiments; 1978-1995*.
10. Hach Co. 1990. *Hach Company Catalog: Production Analysis Ames, IA*.
11. U.S. EPA 1993 *Test Methods for Evaluating Solid Waste; Physical and Chemical Methods*, 3rd Ed. SW846 Washington, D.C.
12. Bouwer, E.J., B.E. Rittman, and P.L. McCarty. 1981 Anaerobic Degradation of Halogenated 1-and 2-Carbon Organic Compounds. *Environ. Sci. Technology*. 15(5): 596-599.
13. Bouwer, E.J., 1994 *Bioremediation of Chlorinated Solvents Using Alternate Electron Aceptors*. In: Norris, R.D., R.E. Hinchee, R. Brown, P.L. McCarty, L. Semprint, J.T. Wilson,

DH. Kampbell, M. Reinhard, E.J. Bower, R.C. Borden, J.M. Vogel, J.M. Thomas, and C.H. Ward, Eds. Handbook of Bioremediation, Boca Raton, Fl.: Lewis Publishers.

14. Vogel, T.M., 1994. Natural Attenuation of Chlorinated Solvents. In: Norris, R.D., R.E. Hinchee, R. Brown, P.L. McCarty, L. Semprini, J.T. Wilson, D.H. Kampbell, M. Reinhard, E.J. Bower, R.C. Borden, T.M. Vogel, J.M. Thomas, and C.H. Ward, eds. Handbook of Bioremediation. Boca Raton, Fl: Lewis Publishers

15. Vogel, T.M and P.L. McCarty. 1995 Biotransformation of Tetrachloroethylene to Trichloroethylene, dichloroethylene, vinyl chloride, and carbon dioxide under methanogenic conditions. Appl. Environ. Microbiol. 49(5):1080-1083.

16. Wiedemeier, T.H., M.A. Swanson, J.T. Wilson, D.H. Kampbell, R.N. Miller, and J.E. Hansen. 1996 Approximation of Biodegradation Rate Constants for Monoaromatic Hydrocarbons in Groundwater. Ground Water Monitoring and Remediation.

17. Wiedemeier, T.H., M.A. Swanson, D.E. Moutoux, J.T. Wilson, D.H. Kampbell, J.E. Hansen, P. Haas, and F.H. Chapelle. 1996. Technical Protocol for Natural Attenuation of Chlorinated Solvents in Groundwater. San Antonio, Tx: U.S. Air Force Center for Environmental Excellence.

16. Phillips, J.B., M. Hindawi, A. Phillips, R.V. Bailey. Engineering Development Institute. Viewing Bioremediation from a Multimedia Perspective. Pollution Engineering, Jan. 98. p.33-35.

17. Schaffner, Jr., P.G., E. Hawkins, J. Wieck. A Look at Degradation of CAHs. Soil and Groundwater Cleanup, May 1996. p.20-31.

18. U.S.EPA Draft Region 4. Suggested Practices for Evaluation of a Site for Natural Attenuation (Biological Degradation) of Chlorinated Solvents. November, 1997.

19. Murry, W.D., and M. Richardson. 1993. Progress Toward the Biological Treatment of C1 and C2 Halogenated Hydrocarbons. Crit. Rev. Environ. Sci. Technology 23(3):195-217

20. McCarty, P.L., and L. Semprini. 1994 Groundwater treatment for chlorinated solvents, In: Norris, R.D., R.E. Hinchee, R. Brown, P.L. McCarty, L. Semprini, J.T. Wilson, D.H. Kampbell, M. Reinhard, E.J. Bower, R.C. Borden, T.M. Vogel, J.M. Thomas, and C.H. Ward, eds. Handbook of bioremediation. Boca Raton, Fl: Lewis Publishers.

21. Linn, W., P.G. Considerations for Assessment of Dry Cleaning Solvent Contaminated Sites-Draft, November 1998. Florida Department of Environmental Protection.

22. Semprini, L., Peter K. Kitanidis, Don H. Kampbell, John T. Wilson. Anaerobic Transformation of Chlorinated Aliphatic Hydrocarbons in a Sand Aquifer Based on Spatial Chemical Distributions. Water Resources Research, 31(4):1051-1062. April 1995.

23. RTDF. Natural Attenuation of Chlorinated Solvents in Groundwater: Principals and Practices. Draft Version 3.0, August 1997.
24. Domenico, RA. 1987. An Analytical Model for Multidimensional Transport of a Decaying Contaminant Species. *J. Hydro.* 91:49-58.
25. Buscheck, T.E.G., and C.M Alacanter. 1995. Regression Techniques and Analytical Solutions to Demonstrate Intrinsic Bioremediation. In: Proceeding of the 1995 Battelle International Conference on In-Situ and On-Site Bioremediation. April
26. Bear, J. 1979. *Hydraulics of Groundwater.* New York, NY McGraw-Hill
27. EPA Seminars. "Monitored Natural Attenuation for Groundwater." Office of Research and Development, U.S. Environmental Protection Agency, Washington, DC. September, 1998.
28. Interstate Technology Regulatory Cooperation (ITRC); Workshop Notes, November, 1998.
29. Cox, E.E., L.Lehmick, E. Edwards, R. Mechaber, B. Su and D.W. Major. 1997. Field and Laboratory Evidence of Sequential Anaerobic-Cometabolic Biodegradation of Chlorinated Solvents. In: *In Situ and On-Site Bioremediation: Vol. 3.* Battelle Press, p. 203.

**APPENDIX A**  
**GLOSSARY OF TERMS**

## **APPENDIX A GLOSSARY OF TERMS**

Parts of this glossary have been adapted from National Research Council (1993) .7

**Abiotic** - Occurring without the involvement of microorganisms.

**Advection** - Transport of molecules dissolved in water along the groundwater flow path at an average expected velocity.

**Aerobic** - Environmental conditions where oxygen is present.

**Aerobic Respiration** - Process whereby microorganisms use oxygen as an electron acceptor to generate energy.

**Aliphatic Hydrocarbon** - A compound built from carbon and hydrogen atoms joined in a linear chain. Petroleum products are composed primarily of aliphatic hydrocarbons.

**Anaerobic** - Environmental conditions where oxygen is absent.

**Anaerobic Respiration** - Process whereby microorganisms use a chemical other than oxygen as an electron acceptor. Common "substitutes" for oxygen are nitrate, sulfate, iron, carbon dioxide, and other organic compounds (fermentation).

**Aquifer** - An underground geological formation that stores groundwater.

**Bacterium** - A single cell organism of microscope size. Bacteria are ubiquitous in the environment, inhabiting water, soil, organic matter and the bodies of plants and animals.

**Biochemical** - Produced by, or involving chemical reactions of living organisms.

**Biodegradation** - Biologically mediated conversion of one compound to another.

**Biomass** - Total mass of microorganisms present in a given amount of water or soil

**Bioremediation** - Use of microorganisms to control and destroy contaminants.

**Biotransformation** - Microbiologically catalyzed transformation of a chemical to some other product.

**Chlorinated Solvent** - A hydrocarbon in which chlorine atoms substitute for one or more hydrogen atoms in the compound's structure. Chlorinated solvents commonly are used for grease removal in manufacturing and dry cleaning, and other operations.

**Co-metabolism** - A reaction in which microbes transform a contaminant even though the contaminant cannot serve as an energy source for the organisms. To degrade the contaminant, the microbes require the presence of other compounds (primary substrates) that can support their growth.

**Degradation** - Destruction of a compound through biological or abiotic reactions.

**Dechlorination** - The removal of chlorine atoms from a compound.

**Desorption** - Opposite of sorption; the release of chemicals attached to solid surfaces.

**Diffusion** - Dispersive process that results from the movement of molecules along a concentration gradient. Molecules move from areas of high concentration to low concentration.

**Dilution** - The combined processes of advection and dispersion result in a net dilution of the molecules in the groundwater.

**Dispersion** - The spreading of molecules along and away from the expected groundwater flow path during advection as a result of mixing of groundwater in individual pores and channels.

**Electron** - A negatively charged subatomic particle that may be transferred between chemical species in chemical reactions. Every chemical molecule contains electrons and protons (positively charged particles).

**Electron Acceptor** - Compound that gains electrons (and therefore is reduced) in oxidation-reduction reactions that are essential for the growth of microorganisms. Common electron acceptors are oxygen, nitrate, sulfate, iron and carbon dioxide. Highly chlorinated solvents (e.g., TCE) can act as electron acceptors.

**Electron Donor** - Compound that loses electrons (and therefore is oxidized) in oxidation-reduction reactions that are essential for growth of microorganisms. In bioremediation organic compounds serve as electron donors. Less chlorinated solvents (e.g., VC) can act as electron donors.

**Geochemical** - produced by, or involving non-biochemical reactions of the subsurface.

**Growth Substrate** - an organic compound upon which a bacteria can grow, usually as a sole carbon and energy source.

**Hydraulic Conductivity** - A measure of the rate at which water moves through a unit area of the subsurface under a unit hydraulic gradient.

**Hydraulic Gradient** - change in head (i.e., water pressure) per unit distance in a given direction, typically in the principal flow direction.

**Inorganic Compound** - A chemical that is not based on covalent carbon bonds. Important examples are metals, nutrients such as nitrogen and phosphorus, minerals, and carbon dioxide.

**Intrinsic Bioremediation** - A type of in situ bioremediation that uses the innate capabilities of naturally occurring microbes to degrade contaminants without taking any engineering steps to enhance the process.

**Intrinsic Remediation** - In situ remediation that uses naturally occurring processes to degrade or remove contaminants without using engineering steps to enhance the process.

**Metabolic Intermediate** - A chemical produced by one step in a multistep biotransformation.

**Metabolism** - The chemical reactions in living cells that convert food sources to energy and new cell mass.

**Methanogen** - A microorganism that exists in aerobic or anaerobic environments and produces methane as the end product of its metabolism. Methanogens use carbon dioxide or simple carbon compounds such as methanol as an electron acceptor.

**Microcosm** - A laboratory vessel set up to resemble as closely as possible the conditions of a natural environment.

**Microorganism** - An organism of microscopic or submicroscopic size. Bacteria are microorganisms.

**Mineralization** - The complete degradation of an organic chemical to carbon dioxide, water, and in some cases inorganic ions.

**Natural Attenuation** - Naturally-occurring processes in soil and groundwater environments that act without human intervention to reduce the mass, toxicity, mobility, volume or concentration of contaminants in those media. **Natural attenuation is the combined effect of dispersion, dilution, volatilization, sorption, chemical reactions**

**with subsurface materials, and biodegradation of dissolved contaminants in the groundwater and soil.**

**Nonaqueous Phase Liquids (NAPLs)** - An organic liquid that is maintained as a separate phase from water.

**Oxidization** - Loss of electrons from a compound, such as an organic contaminant. The oxidation can supply energy that microorganisms use for growth. Often, but not always, oxidation results in the addition of an oxygen atom and/or the loss of a hydrogen atom.

**Oxygenase** - An enzyme that introduces oxygen into an organic molecule.

**Plume** - A zone of dissolved contaminants. A plume usually originates from a source and extends for some distance in the direction of ground water flow.

**Primary Substrates** - The electron donor and electron acceptor that are essential to ensure the growth of microorganisms. These compounds can be viewed as analogous to the food and oxygen that are required for human growth.

**Reduction** - Transfer of electrons to a compound such as oxygen. It occurs when another compound is oxidized.

**Reductive Dechlorination** - The removal of chlorine atoms from an organic compound and their replacement with hydrogen atoms (same as reductive dehalogenation).

**Reductive Dehalogenation** - A variation on biodegradation in which microbially catalyzed reactions cause the replacement of a halogen atom (e.g., chlorine) on an organic compound with a hydrogen atom. The reactions result in the net addition of two electrons to the organic compound.

**Saturated Zone** - Subsurface environments where pore spaces are filled with water.

**Site Conceptual Model** - A hypothesis about how releases occurred, the current state of the source, an idealized geochemical site type, and current plume characteristics (plume stability).

**Sorption** - Attachment of a substance on the surface of a solid by physical or chemical attraction.



**Stabilization** - Process whereby chemical molecules become chemically bound by a stabilizing agent (e.g., clay, humic materials), reducing the mobility of the molecule in groundwater.

**Substrate** - A compound that microorganisms can use in the chemical reactions catalyzed by their enzymes.

**Sulfate Reducer** - A microorganism that exists in anaerobic environments and reduces sulfate to hydrogen sulfide.

**Unsaturated Zone** - Soil above the water table, where pores are partially or largely filled with air.

**Vadose Zone** - See "Unsaturated zone."

**Volatilization** - Transfer of a chemical from the liquid to the gas phase (as in evaporation).