White Paper

 Development of Surface Water Screening Levels for PFOA and PFOS Based on the Protection of Human Health Using Probabilistic Risk Assessment

 Prepared for the District and Business Support Program Florida Department of Environmental Protection

Leah Stuchal, Ph.D. and Stephen M. Roberts, Ph.D.

 Center for Environmental & Human Toxicology University of Florida

April 2020

 This white paper develops surface water screening levels for perfluorooctanoic acid human health using probabilistic risk assessment (PRA). These screening levels are based on fish and shellfish ingestion pathways. PFOA and PFOS are manmade chemicals that belong to a group of thousands of chemicals known as perfluoroalkyl substances (PFAS). PFASs are carpets, leather, textiles, furniture, and packaging materials. They are also used in fire-fighting foam and are added to aviation fluids to decrease flammability. PFOA and PFOS degrade USEPA, 2016b). The PFOA and PFOS present in surface water bioconcentrates and bioaccumulates into fish and shellfish that are consumed by local populations. (PFOA; CAS# 335-67-1) and perfluorooctane sulfonate (PFOS; CAS# 1763-23-1) protective of water- and lipid-resistant. They are used as waterproofing and stain-resistant coatings for slowly and are very persistent in the environment and the human body (USEPA, 2016a;

 bioaccumulates into fish and shellfish that are consumed by local populations. The following sections describe the technical basis for the proposed surface water screening levels.

Equation and assumptions

 We calculated surface water screening levels protective of fish and shellfish consumption using a modified equation from the U.S. Environmental Protection Agency (USEPA) for the calculation of fish consumption limits based on concentrations of contaminants in fish tissue (USEPA, 2000a). The equation for non-carcinogens was used, modified by removing the drinking water intake component.

$$
SWSL (\mu g/L) = \frac{RfD \times RSC \times BW \times CF}{\sum_{i=2}^{4} (Fl_i \times BAF_i)}
$$

Where:

 $SWSL = surface$ water screening level (μ g/L)

RfD = oral reference dose (mg/kg-d)

RSC = relative source contribution

BW = body weight (kg)

 $CF = correction factor, 1000 \mu g/mg$

 FI_i = freshwater and estuarine finfish and shellfish consumption rate for aquatic trophic levels 2, 3, and 4 (kg/d)

 BAF_i = bioaccumulation factor for aquatic trophic levels (TLs) 2, 3, and 4 (L/kg)

 $\sum_{i=2}^{4}$ = summation of values for aquatic TLs, where the letter i stands for the TLs, starting with TL2 and continuing to TL4

 consumption rate (fish consumption rate) were chosen as distributions. Point values were selected for the other exposure parameters. This is identical to the PRA method proposed for the surface water standards in Chapter 62-302, F.A.C. (FDEP, 2016). The point value parameters are listed in Table 1 and the distributions for fish ingestion are provided in Table 2. Body weight was defined as a lognormal distribution with a mean of 79.96 kg and a standard deviation of 20.73 kg (USEPA, 2011). Figures showing the distributions for body weight and For the PRA, body weight and freshwater and estuarine finfish and shellfish fish ingestion are included in Appendix A.

Parameter	PFOA	PFOS	Source
Reference dose (mg/kg-d)	2E-05	2E-05	USEPA, 2016a USEPA, 2016b
Relative source contribution	0.6	0.6	CEHT, 2020
Bioaccumulation factor $TL2$ (L/kg)	35	937	See section on bioaccumulation factor
Bioaccumulation factor $TL3$ (L/kg)	71	2959	See section on bioaccumulation factor
Bioaccumulation factor TL4 (L/kg)	161	6304	See section on bioaccumulation factor

 Table 2 – Fish ingestion lognormal distributions used in the PRA for the derivation of surface water screening levels for PFOA and PFOS

 The fish ingestion distributions were derived from USEPA, 2014, Appendix E, Tables E-13, E-14, and E-15; NA – not applicable. This statistic was not used to define the distribution.

Reference Dose

 The USEPA has developed reference doses for PFOA and PFOS in order to create drinking water Health Advisory Levels for these compounds. FDEP has used these reference doses for the calculation of alternative groundwater cleanup target levels (GCTLs) and soil cleanup target levels (SCTLs) for PFOA and PFOS (See letters to the FDEP dated April 16, 2018 and August 16, 2018 for details regarding the derivation of those screening levels). For consistency, the same RfD values are used in the surface water calculation, i.e., an oral reference dose (RfD) of 2E-05 mg/kg-d for both PFOA and PFOS.

 We are aware that there is a lack of consistency among federal and state agencies in the derivation of safe limits for oral exposure to these substances. The Agency for Toxic Substances and Disease Registry (ATSDR) released a draft toxicity profile for PFAS, including PFOA and PFOS. The proposed Minimal Risk Levels (MRLs; analogous to RfDs) are an order of magnitude lower than the USEPA RfDs. This draft document received extensive public comment and has not yet been finalized. Additionally, North Carolina, Texas, Maine, Minnesota, and New Jersey have developed toxicity values for PFOA and PFOS based on

 differing endpoints and/or uncertainty factors. This results in different toxicity values than were proposed by the USEPA and ATSDR. California has also derived slope factors for PFOA and PFOS based on the development of pancreatic and liver tumors in male rats (CalEPA, 2019). The potential toxicity of PFOA and PFOS is a subject of active research, and the data available are rapidly evolving. Thus, while the USEPA RfD values are used for the surface water information develops. Use of toxicity values developed based upon other endpoints, including cancer, instead of the USEPA RfDs will result in different screening level estimates that may be lower than those calculated here. screening levels proposed here, we recommend re-visiting these screening levels as new

Body Weight

 The Exposure Factors Handbook recommends using the body weight distributions calculated by Portier et al., (2007) for probabilistic risk assessment. For this analysis, body weight was defined as a lognormal distribution with a mean of 79.96 kg and a standard deviation of 20.73 kg (USEPA, 2011). This distribution represents the National Health and Nutrition Examination Survey (NHANES) IV estimated body weights for 18 to 65-year-old males and females. It was not truncated for the risk assessment. This body weight distribution was also used in the FDEP (2016) technical support document for the derivation of surface water standards.

Relative Source Contribution

 This assessment uses the USEPA relative source contribution (RSC) values of 0.6 (60%) for both PFOA and PFOS. These chemical-specific RSCs for PFOA and PFOS were derived using the USEPA Exposure Decision Tree methodology (USEPA, 2000b), as explained in the companion white paper, "*Determination of Relative Source Contribution Values for* Perfluorooctanoic Acid (PFOA) and Perfluorooctane Sulfonate (PFOS) in Support of *Development of Florida Surface Water Screening Levels*" (CEHT, April 2020). As described in this document, there are several potential sources for PFOA and PFOS identified in the literature, of which drinking water, diet, consumer products, and indoor air and dust may be important. exposure from drinking water consistent with current FDEP screening levels, dietary exposure, and potential exposure to other sources combined. From this analysis, 60% of the intake corresponding to the RfDs for PFOA and PFOS were allocated to surface water exposure in the RSC values were derived using the percentage method, taking into account form of consumption of fish and shellfish.

Finfish and Shellfish Consumption Rate

 No current Florida-specific fish consumption studies are available. The Degner et al. (1994) fish consumption study in Florida was used previously to develop fish consumption rates for Chapter 62-302, F.A.C. It also includes Florida-specific data on shellfish consumption for the general population. However, it is dated and may not represent current fish consumption rates. Therefore, we used NHANES 2003-2010 fish consumption data. The NHANES data are presented based on age, gender, and geographic region. The USEPA document summarizing the NHANES fish and shellfish consumption data presents several consumer categories that may be relevant to fish consumption in Florida (USEPA, 2014). However, none are specific to Florida. Based on the differences in fish and shellfish consumption rates for each geographic region in the U.S., these national data may not be appropriate for Florida. Regional data applicable to Florida include the South, Gulf of Mexico coastal counties, and Atlantic coastal counties. The FDEP determined that the combination of these regions provide the best estimate for a fish consumption rate for Floridians. They calculated that 44.8% of the Florida population resides within the Atlantic coastal counties, 31.6% reside in the Gulf of Mexico coastal counties, and 23.6% reside in the South geographic region (FDEP, 2016).

 estimates of fish consumption by trophic level for each region. We used fit statistics to describe lognormal distributions for the consumption rates in Tables E-13, E-14, and E-15 of that document (Table 2). These statistics represent the total freshwater and estuarine finfish and shellfish consumption rates for adults. The USEPA document summarizing the NHANES fish consumption data also includes

Bioaccumulation Factor

 Bioaccumulation factors (BAFs) for PFOA and PFOS from the literature are listed in Appendix Tables B1 and B2. Studies that included data on bioaccumulation in the muscle tissue (e.g., fillet) were utilized to calculate the freshwater BAFs. These studies include fish not present in Florida (e.g., rainbow trout) and fish not usually consumed (e.g., minnows, whitebait). The BAFs for these fish were used to calculate a freshwater BAF for PFOA and PFOS because bioaccumulation data in fish and shellfish are limited. By including all of the data available, it provides a better estimate of the BAF.

 Bioaccumulation factors for the derivation of human health surface water criteria were 2016c). Based on the USEPA proposed framework, we utilized field BAFs to calculate baseline BAFs for PFOA and PFOS. Field BAFs are the preferred source for calculating BAFs for nonionic organic chemicals. Typically, in this methodology, a baseline BAF is calculated based on the field BAF, the concentration of particulate organic carbon (POC) in the water, the water partition coefficient (Kow), and the fraction of finfish and shellfish tissue that is lipid. However, for the purposes of this assessment, the field BAFs were used as the baseline BAFs. The reasoning for this includes: calculated based on a modified version of the USEPA framework for deriving BAFs (USEPA, concentration of dissolved organic carbon (DOC) in the water, the chemical-specific n-octanol-

- 1. The POC and DOC were not known for the majority of the BAF studies. Calculation of the fraction of chemical in water that is freely dissolved would require assumptions regarding the amount of dissolved and particulate carbon. Although national averages may be used as defaults, the majority of studies took place outside the United States and default POC and DOC values for these countries are unknown.
- 2. The K_{ow} has not been measured for PFOA and PFOS. Calculation of a baseline BAF would require a K_{ow} based on physical/chemical property estimation software (e.g., EPI Suite).
- 3. Unlike most non-ionic organics, PFAS are not distributed to the lipid. Therefore, use of a lipid adjustment to derive a baseline BAF is inappropriate for PFAS chemicals.

 Bioaccumulation factors were derived for each trophic level (TL). To calculate a BAF, the fish and shellfish from the bioaccumulation studies were assigned to trophic levels (Table B3). A bioaccumulation factor was calculated for each trophic level for both PFOA and PFOS. Individual field BAFs were combined as the geometric mean for each species. The baseline TL- specific BAF was calculated as the geometric mean of all species geometric means (Table 3). These calculations are presented in Tables B4 though B9. The Minnesota study (MPCA, 2007b) combined bluegill and white bass in their river bioaccumulation study. Because the BAFs were listed as geometric means (MPCA, 2007b), they were retained in the assessment.

Chemical	Trophic Level	BAF
		35
PFOA		
		161
		937
PFOS		2959
		6304

 Table 3 – Trophic level 2, 3, and 4 geometric mean bioaccumulation factors for PFOA and PFOS

Method for the probabilistic risk assessment

 PFOA and PFOS surface water screening levels were calculated using PRA. In this using 100,000 iterations of a Monte Carlo analysis in Crystal Ball software (Version 11.1) with a seed of 123457. Each iteration represents a hypothetical person in the population. For each iteration, the software chose a body weight from the distribution. Then, a region was chosen based on the percentage of Floridians who live in each area. Once the region was identified, the software chose a region-specific fish consumption rate for trophic levels 2, 3, and 4 (Table 2). The fish consumption rates for each trophic level were multiplied by their respective BAFs (Table 3) before being summed. There was no correlation between the fish consumption rates for the three TLs. We could not locate any data suggesting that a high-end consumer of fish and shellfish in TL2 would also be a high-end consumer of fish and shellfish in the other TLs. water concentrations equivalent to a hazard index of 1 for each iteration. The PFOA and PFOS screening levels were set at the $10th$ percentile of this distribution. To check these values, the equation was rearranged and solved to ensure that the hazard index of 1 was not exceeded at the $90th$ percentile (Chapter 62-780, F.A.C.). analysis, body weight and fish consumption were defined as distributions and the other parameters were entered as point values. The surface water screening levels were derived in Using the equation provided in this document, the software generated a distribution of surface

Screening levels

 Surface water screening levels for PFOA and PFOS were calculated using the equations and We also calculated the hazard index for the screening levels to insure it was below 1 at the $90th$ percentile. The hazard index for PFOA at a surface water screening level of 0.5 μ g/L is 1 and the hazard index for PFOS at a screening level of 0.01 μ g/L is 0.8 at the 90th percentile. The screening level for PFOS at a hazard index of 1 is 0.012 μ g/L. This was rounded to 1 significant figure, which decreased the hazard index at the 90th percentile. The Crystal Ball output for the presented in Appendix C. assumptions described in this document. The surface water screening level for PFOA is 0.5 μ g/L and for PFOS is 0.01 μ g/L (Table 4). The distributions are presented in Figures 1 and 2. surface water screening level distributions and the distributions for the hazard index are

Table 4 – Surface water screening levels for freshwater and estuarine

finfish and shellfish for PFOA and PFOS				
Surface Water Screening	PFOA	PFOS		
Levels $(\mu g/L)$				
Freshwater and estuarine finfish and shellfish	0.5	0.01		

Screening levels were rounded to one significant figure

Figure 1 – Surface water screening level distribution for PFOA

Figure 2 – Surface water screening level distribution for PFOS

 The USEPA drinking water Health Advisory Levels (HALs) for PFOA and PFOS are each 0.07 µg/L. The USEPA recommends that the combined concentrations of PFOA and PFOS in drinking water be compared with this limit, based upon an assumption that their effects are additive. The rationale for this assumption is that their RfDs are derived for the same toxic endpoint (developmental effects) and that, although the mode of action for these effects has not been established, it is likely to be the same for these closely related chemicals. The fact that their individual Health Advisory Levels are identical makes it relatively straightforward to implement this recommendation. While the same argument could be made that the surface water screening levels for PFOA and PFOS should also address combined effects, this is more difficult because of the large difference in their values, approximately an order of magnitude. Picking the lower, higher, or average of these values for comparison with combined PFOA and PFOS concentrations could result in gross over- or underestimation of risk, depending on the individual PFOA and PFOS concentrations. As a practical matter, comparison of PFOA and PFOS concentrations in surface water with the screening levels should be made individually.

Alaska 0.07* 0.07*

Surface water screening levels in Florida and other states

* - Concentrations of PFOA and PFOS are summed before being compared to the criterion.

 Minnesota has also developed freshwater surface water criteria based on fish consumption for the protection of human health. These criteria are based on site-specific bioaccumulation factors. For PFOA, the Minnesota surface water criteria include 1.6 µg/L for lakes and 2.7 µg/L for rivers (MPCA, 2017; Table 5). These criteria are higher than our proposed screening level of 0.5 μ g/L. The difference in values is due to the use of a higher oral reference dose (1.4E-04 mg/kg-d) and slightly lower bioaccumulation factor (40 L/kg for lakes and 24 L/kg for rivers). Recently, the Minnesota Department of Health (MDOH) updated their reference doses for PFOA and PFOS (MDOH, 2019a; MDOH, 2019b). The updated reference dose for PFOA is 1.8E-05 mg/kg-d (MDOH, 2019a). Using this reference dose in their surface water equation would decrease the Minnesota criterion by approximately one order of magnitude. These updated values would be slightly lower than our proposed screening level of 0.5 µg/L.

 The screening level for PFOS is lower than PFOA due to the large bioaccumulation factor for PFOS. For PFOS, the Minnesota surface water criteria include 0.006 µg/L for lakes 0.007 µg/L for rivers (MPCA, 2017). Our proposed PFOS screening level of 0.01 µg/L is similar to these two criteria. This is due to the use of a similar reference dose (8E-05 mg/kg-d) and 2010b). The MDOH updated reference dose for PFOS is 3.1E-06 mg/kg-d (MDOH, 2019b). Use of this reference dose would lower the PFOS criteria to less than $0.001 \mu g/L$, which is an order of magnitude below our proposed screening level. bioaccumulation factors (6,087 L/kg for lakes and 3,877 for rivers) (MPCA, 2010a; MPCA,

 The Michigan Department of Environmental Quality (MDEQ) criteria for PFOA and PFOS are human health-based non-cancer values for non-drinking surface water sources. They were derived based on Michigan Rule 57 for toxic substances (MDEQ, 2020; Table 5). Their surface water screening level for PFOA is an order of magnitude greater than the screening level of 0.5 µg/L proposed in this document. The Michigan surface water screening level for PFOS is equivalent to our proposed PFOS surface water value of 0.01 µg/L. The Alaska Department of Environmental Conservation uses a criterion of 0.07 µg/L for PFAS in surface water used as drinking water (ADEC, 2019; Table 5). The criterion includes the sum of PFOA and PFOS concentrations. It is based on the USEPA drinking water HAL.

References:

- ADEC (2019) *Technical Memorandum Action Levels for PFAS in Water and Guidance on* Sampling Groundwater and Drinking Water. Alaska Department of Environmental Conservation, Division of Spill Prevention and Response Contaminated Sites Program and Division of Environmental Health Drinking Water Program, Juneau, Alaska.
- CalEPA (2019) *Notification Level Recommendations: Perfluorooctanoic Acid and* Perfluorooctane Sulfonate in Drinking Water. Pesticide and Environmental Toxicology Branch, Office of Environmental Health Hazard Assessment, California Environmental Protection Agency.
- CEHT (2020) *Determination of Relative Source Contribution Values for Perfluorooctanoic Acid (PFOA) and Perfluorooctane Sulfonate (PFOS) in Support of Development of Florida Surface Water Screening Levels*. Center for Environmental and Human Toxicology, University of Florida, Gainesville, Florida.
- Degner, RL, Adams, CM, Moss, SD, Mack, SK (1994) *Per Capita Fish and Shellfish Consumption in Florida.* Florida Agricultural Market Research Center, Food and Resource Economics Department, Institute of Food and Agricultural Sciences, University of Florida, Gainesville, Florida.
- Fang, S, Chen, X, Zhao, S, Zhang, Y, Jiang, W, Yang, L, Zhu, L (2014) Trophic magnification and isomer fractionation of perfluoroalkyl substances in the food web of Taihu Lake, China. *Environmental Science and Technology* 48: 2173-2182.
- Fang, S, Zhang, Y, Zhao, S, Qiang, L, Chen, M, and Zhu, L (2016) Bioaccumulation of Perfluoroalkyl acids including the isomers of perfluorooctane sulfonate in carp (*Cyprinus carpio)* in sediment/water microcosm. *Environmental Toxicology and Chemistry* 35(12): 3005-3013.
- FDEP (2016) *Technical Support Document: Derivation of Human Health-Based Criteria and Risk Impact Statement.* Florida Department of Environmental Protection, Division of Environmental Assessment and Restoration, Tallahassee, Florida.
- Martin, JW, Mabury, SA, Solomon, KR, Muir, DCG (2003) Bioconcentration and tissue *Toxicology and Chemistry* 22(1): 196-204. distribution of perfluorinated acids in rainbow trout (*Oncorhynchus mykiss)*. *Environmental*
- MPCA (2007a) *Surface Water Quality Criterion for Perfluorooctanoic Acid.* Minnesota Pollution Control Agency, St. Paul, Minnesota.
- MPCA (2007b) *Surface Water Quality Criterion for Perfluorooctane Sulfonic Acid.* Minnesota Pollution Control Agency, St. Paul, Minnesota.
- MPCA (2010a) *Minnesota Pollution Control Agency, Aquatic Life Criteria and Water Quality Standards for [PFOS].* Minnesota Pollution Control Agency, St. Paul, Minnesota. <https://www.pca.state.mn.us/sites/default/files/pfos-lakecalhoun.pdf>
- MPCA (2010b) *Mississippi River Pool 2 Intensive Study of Perfluorochemicals in Fish and Water: 2009.* Minnesota Pollution Control Agency, St. Paul, Minnesota.
- MPCA (2017) *Human Health-based Water Quality Standards Technical Support Document,* Water Quality Standard Amendments - Minn. R. chs. 7050 and 7052 [Final]. Minnesota Pollution Control Agency, St. Paul, Minnesota.
- MDEQ (2020) *Rule 57 Surface Water Quality Values, Surface Water Assessment Section, Michigan EGLE.* Michigan Department of Environmental Quality. February 1, 2020. https://www.michigan.gov/egle/0,9429,7-135-3313_3681_3686_3728-11383--,00.html
- MDOH (2019a) *Toxicological Summary for: Perfluorooctanoate.* Minnesota Department of Health, Health Based Guidance for Water, Health Risk Assessment Unit, Environmental Health Division, St. Paul, MN.
- MDOH (2019b) *Toxicological Summary for: Perfluorooctane sulfonate.* Minnesota Department of Health, Health Based Guidance for Water, Health Risk Assessment Unit, Environmental Health Division, St. Paul, MN.
- Naile, JE, Khim, JS, Hong, S, Park, J, Kwon, B, Ryu, JS, Hwang, JH, Jones, PD, Giesy, JP (2013) Distributions and bioconcentration characteristics of perfluorinated compounds in environmental samples collected from the west coast of Korea. *Chemosphere* 90: 387- 394.
- USEPA (2000a) *Guidance for Assessing Chemical Contaminant Data for Use in Fish Advisories, Volume 2, Risk Assessment and Fish Consumption Limits, Third Edition.* United States Environmental Protection Agency, Office of Water, Washington, DC.
- USEPA (2000b) Methodology for Deriving Ambient Water Quality Criteria for the Protection of *Human Health*. EPA-822-B-00-004. U.S. Environmental Protection Agency, Office of Water, Office of Science and Technology, Washington, DC.
- USEPA (2011) *Exposure Factors Handbook: 2011 Edition*. United States Environmental Protection Agency, National Center for Environmental Assessment, Office of Research and Development, Washington, DC.
- USEPA (2014) *Estimated Fish Consumption Rates for the U.S. Population and Selected Subpopulations (NHANES 2003-2010).* United States Environmental Protection Agency, Washington, DC.
- USEPA (2016a) *Drinking Water Health Advisory for Perfluorooctanoic Acid (PFOA)*. United States Environmental Protection Agency, Office of Water, Washington, DC.
- USEPA (2016b) *Drinking Water Health Advisory for Perfluorooctane Sulfonate (PFOS).* United States Environmental Protection Agency, Office of Water, Washington, DC.
- USEPA (2016c) *Development of National Bioaccumulation Factors: Supplemental Information for EPA's 2015 Human Health Criteria Update.* United States Environmental Protection Agency, Office of Water, Office of Science and Technology, Washington, DC.
- Zhou, Z, Shi, Y, Li, W, Xu, L, Cai, Y (2012) Perfluorinated compounds in surface water and potential risk. *Bulletin of Environmental Contamination and Toxicology* 89: 519-524. organisms from Baiyangdian Lake in North China: Source profiles, bioaccumulation and

Appendix A

Crystal Ball Report - Assumptions

Assumptions

Assumption: Body weight · PFOA Value

Assumption: Body weight · PFOS Value

Assumption: Fish consump TL2 Atlantic · PFOA Value

Assumption: Fish consump TL2 Atlantic · PFOS Value

Assumption: Fish consump TL2 Gulf · PFOA Value

Assumption: Fish consump TL2 Gulf · PFOS Value

Assumption: Fish consump TL2 South · PFOA Value

Assumption: Fish consump TL2 South · PFOS Value

Assumption: Fish consump TL3 Atlantic · PFOA Value

Assumption: Fish consump TL3 Atlantic · PFOS Value

 0.0

Assumption: Fish consump TL3 Gulf · PFOA Value

Lognormal distribution with parameters:

Assumption: Fish consump TL3 Gulf · PFOS Value

Assumption: Fish consump TL3 South · PFOA Value

Assumption: Fish consump TL3 South · PFOS Value

Assumption: Fish consump TL4 Atlantic · PFOA Value

Assumption: Fish consump TL4 Atlantic · PFOS Value

Assumption: Fish consump TL4 Gulf · PFOA Value

Assumption: Fish consump TL4 Gulf · PFOS Value

Assumption: Fish consump TL4 South · PFOA Value

Assumption: Fish consump TL4 South · PFOS Value

 0.0 Mean 2.8 15.8

Appendix B

Species	Place	Fresh or Marine	Exposure	Tissue	BAF	Study
Common carp	laboratory	Fresh	static 28d, 28d flow through depuration	muscle	3.85	Fang et al, 2016
Minnow	Faihu Lake, China	Fresh	wild caught	muscle	112.5	Fang et al., 2014
Silver carp	Taihu Lake, China	Fresh	wild caught	muscle	11.8	Fang et al., 2014
Whitebait	Taihu Lake, China	Fresh	wild caught	muscle	147	Fang et al., 2014
Crucian carp	Taihu Lake, China	Fresh	wild caught	muscle	81	Fang et al., 2014
Lake saury	「aihu Lake, China	Fresh	wild caught	muscle	284	Fang et al., 2014
Common carp	Taihu Lake, China	Fresh	wild caught	muscle	177	Fang et al., 2014
Mongolian culter	Гаіhu Lake, China	Fresh	wild caught	muscle	161	Fang et al., 2014
Mud fish	Faihu Lake, China	Fresh	wild caught	muscle	163	Fang et al., 2014
Chinese bitterling	Гаіhu Lake, China	Fresh	wild caught	muscle	87.9	Fang et al., 2014
Goby	Taihu Lake, China	Fresh	wild caught	muscle	37.7	Fang et al., 2014
Common carp	China	Fresh	wild caught	muscle	182	Zhou et al., 2012
White shrimp	Taihu Lake, China	Fresh	wild caught	soft part	12.5	Fang et al., 2014
Pearl mussel	Taihu Lake, China	Fresh	wild caught	soft part	39.7	Fang et al., 2014

Table B1 – Bioaccumulation factors for PFOA in fish fillets and shellfish tissue

		Trophic Level 2	Trophic Level 3	Trophic Level 4	
Fish and shellfish	Scientific Name	Weighting	Weighting	Weighting	Reference
Minnow	Hemiculter leucisculus				FishBase.org
Silver carp	Hypophthalmichthys molitrix			O	USEPA 2014
Whitebait	Reganisalanx brachyrostralis	ი		0	FishBase.org
Crucian carp	Carassius cuvieri	0		0	USEPA 2014
Lake saury	Coilia mystus	0		0	FishBase.org
Common carp	Cyprinus carpio	0		0	USEPA 2014
Mongolian culter	Culter mongolicus	0	0.5	0.5	FishBase.org
Mud fish	Misgurnus anguillicaudatus	0		0	FishBase.org
Chinese bitterling	Rhodeus sinensis	0.5	0.5	0	FishBase.org
Goby	Ctenogobius giurinus	0		0	FishBase.org
White shrimp	Exopalaemon sp.	0.5	0.5	0	USEPA 2014
Pearl mussel	Lamellibranchia sp.		Ω	0	USEPA 2014
Bluegill	Lepomis macrochirus	0		$\mathbf{0}$	FishBase.org
White bass ^a	Morone chrysops	0	∩		USEPA 2014
Taihu Lake shrimp	Macrobrachium nipponense	0.5	0.5	O	USEPA 2014
Freshwater mussel	Lamellibranchia sp.		O	0	USEPA 2014

Table B3 – Trophic level weighting for fish and shellfish in the PFOA and PFOS bioaccumulation studies

a – The white bass value is the geometric mean concentration of *Lepomis macrochirus* and *Morone chrysops*

Species	Scientific name	TI 2 BAF		
Chinese bitterling	Rhodeus sinensis	87.9		
White shrimp	Exopalaemon sp.	12.5		
Pearl mussel	Lamellibranchia sp.	39.7		
	Geometric mean TL2 BAF	35		

Table B4 – Species-specific TL2 BAFs and geometric mean TL2 BAF for PFOA

Species	Scientific name	TL3 BAF
Minnow	Hemiculter leucisculus	112.5
Silver carp	Hypophthalmichthys molitrix	11.8
Whitebait	Reganisalanx brachyrostralis	147
Crucian carp	Carassius cuvieri	81
Lake saury	Coilia mystus	284
Common carp	Cyprinus carpio	50
Mongolian culter	Culter mongolicus	161
Mud fish	Misgurnus anguillicaudatus	163
Chinese bitterling	Rhodeus sinensis	87.9
Goby	Ctenogobius giurinus	37.7
White shrimp	Exopalaemon sp.	12.5
	Geometric mean TL3 BAF	

Table B5 – Species-specific TL3 BAFs and geometric mean TL3 BAF for PFOA

Species	Scientific name	TL2 BAF
Chinese bitterling	Rhodeus sinensis	2861
White shrimp	Exopalaemon sp.	978
Pearl mussel	Lamellibranchia sp.	466
Taihu Lake shrimp	Macrobrachium nipponense	2161
Freshwater mussel	Lamellibranchia sp.	256
	Geometric mean TL2 BAF	937

Table B7 – Species-specific TL2 BAFs and geometric mean TL2 BAF for PFOS

Species	Scientific name	TL3 BAF
Minnow	Hemiculter leucisculus	3212
Silver carp	Hypophthalmichthys molitrix	832
Whitebait	Reganisalanx brachyrostralis	1350
Crucian carp	Carassius cuvieri	6898
Lake saury	Coilia mystus	4401
Common carp	Cyprinus carpio	7433
Mongolian culter	Culter mongolicus	6927
Mud fish	Misgurnus anguillicaudatus	4854
Chinese bitterling	Rhodeus sinensis	2861
Goby	Ctenogobius giurinus	2876
White shrimp	Exopalaemon sp.	978
Bluegill	Lepomis macrochirus	2802
Taihu Lake shrimp	Macrobrachium nipponense	2161
	Geometric mean TL3 BAF	2959

Table B8 – Species-specific TL3 BAFs and geometric mean TL3 BAF for PFOS

Species	Scientific name	TL4 BAF
Mongolian culter	Culter mongolicus	6927
White bass	Morone chrysops	5737
	Geometric mean TL4 BAF	6304

Table B9 – Species-specific TL4 BAFs and geometric mean TL4 BAF for PFOS

Appendix C

Crystal Ball Report - Forecasts

Forecasts

Forecast: SWQC · PFOA

 Entire range is from 0.00 to 55.36 Base case is 1.16 After 100,000 trials, the std. error of the mean is 0.01 Summary:

Forecast: SWQC · PFOA (cont'd)

Forecast: SWQC · PFOS

Summary:

 Entire range is from 0.00 to 1.30 Base case is 0.03 After 100,000 trials, the std. error of the mean is 0.00

Forecast: SWQC · PFOS (cont'd)

Forecast: Hazard index at SWQC PFOA

Summary:

 Entire range is from 8.85E-03 to 1.30E+02 Base case is 4.21E-01 After 100,000 trials, the std. error of the mean is 2.93E-03

Forecast: Hazard index at SWQC PFOA (cont'd)

Forecast: Hazard index at SWQC PFOS

Summary:

 Entire range is from 7.67E-03 to 6.68E+01 Base case is 3.28E-01 After 100,000 trials, the std. error of the mean is 2.17E-03

Forecast: Hazard index at SWQC PFOS (cont'd)

